



HAL
open science

Redox control of chromium in the red soils from China evidenced by Cr stable isotopes

Xiaoquan Qin, Damien Guinoiseau, Zongling Ren, Marc F Benedetti

► **To cite this version:**

Xiaoquan Qin, Damien Guinoiseau, Zongling Ren, Marc F Benedetti. Redox control of chromium in the red soils from China evidenced by Cr stable isotopes. *Journal of Hazardous Materials*, 2023, pp.133406. 10.1016/j.jhazmat.2023.133406 . hal-04375389

HAL Id: hal-04375389

<https://u-paris.hal.science/hal-04375389v1>

Submitted on 5 Jan 2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 *Redox control of chromium in the red soils from China*
2 *evidenced by Cr stable isotopes*

3 *Xiaoquan Qin¹, Damien Guinoiseau², Zongling Ren³ and Marc*
4 *F. Benedetti¹*

5 1 Université Paris Cité – Institut de Physique du globe de Paris, CNRS, F75005 Paris, France

6 2 Université Paris Saclay, CNRS, UMR8148 GEOPS, 91400 Orsay, France

7 3 Department of Soil Science, College of Natural Resources and Environment, South China
8 Agricultural University, Guangzhou 510642, China

9
10 **Highlights**

- 11 • Cr isotope data in soil FMNs is reported for the first time.
12 • FMNs/Fe oxides capture isotopically heavy Cr (VI).
13 • Redox and not land use drives Fe oxide formation and FMNs, affecting Cr mobility.
14 • Cr in red soil profiles mainly results from colluvium mineral weathering.

15 **Abstract**

16 With chromium isotopes, we study the intricate dynamics of adsorption and redox processes in
17 soil ecosystems, focusing on chromium's behaviour, in red soil profiles enriched with iron-manganese
18 nodules (FMNs) in South China. Key findings reveal that the primary geological source of chromium
19 in the red soil profiles is the weathering of colluvium parent minerals. FMNs have higher chromium
20 concentrations (325-1,451 µg/g) compared to surrounding soils (95-247 µg/g) and display stable $\delta^{53}\text{Cr}$
21 values ($0.78 \pm 0.17\text{‰}$), indicating their role as stable chromium repositories, reflecting historical
22 processes. Furthermore, by isolating chromium associated with iron oxides (FeO) and silicate minerals
23 (ReS) within FMNs and surrounding soils using CBD extractions, we show that FeO predominantly
24 carry chromium, particularly in FMNs. The $\delta^{53}\text{Cr}$ values of FeO fractions consistently exhibit heavier

25 signatures than ReS fractions, suggesting the sequestration of isotopically heavy chromium (VI) during
26 Fe oxide precipitation. Fluctuations in soil's redox, rather than land use, play a pivotal role in
27 controlling the precipitation of Fe oxides in surrounding soils and the formation of FMNs, thus
28 influencing chromium mobility. This highlights the significance of these factors when utilizing
29 chromium isotopic techniques for source tracking in soil systems, contributing to our understanding
30 of chromium's behaviour in soil environments.

31 **Keywords**

32 Chromium hazard, Chromium isotopes, red soil, iron-manganese nodules, soil redox, Iron oxides

33 **1. Introduction**

34 In soils, stable Cr is found as trivalent (III) and hexavalent (VI) species with low concentrations
35 (15-100 $\mu\text{g/g}$) in general (Shahid et al., 2017). However, high Cr levels (180-60,000 $\mu\text{g/g}$) can be
36 measured in soils developed over mafic and ultramafic rocks like serpentinite or basalt (Fandeur et al.,
37 2009; Paulukat et al., 2015; Sun et al., 2022). Anthropogenic activities, such as mining, smelting, paper
38 and leather manufacturing, result in $\sim 2.3 \text{ Mg}\cdot\text{yr}^{-1}$ of chromium flux into soils and are the major
39 contributors to soil chromium contamination (Rauch and Pacyna, 2009). Therefore, soil Cr
40 contamination is becoming a widespread concern due to its impact on the human food chain (Ao et al.,
41 2022a; Shahid et al., 2017).

42 Cr is a redox-sensitive element, and manganese oxides are its main natural oxidant in soil. The
43 transformation from rock-derived chromium (III) to more mobile and hazardous hexavalent Cr (CrO_4^{2-} ,
44 HCrO_4^- , $\text{Cr}_2\text{O}_7^{2-}$) has been documented in natural systems like groundwaters (Qin and Wang, 2017).
45 Negatively charged Cr (VI) can be sorbed onto iron oxides or clay minerals, reducing its mobility and
46 risk (Covelo et al., 2008; Dimos et al., 2012; Johnston and Chrysochoou, 2014; Yu et al., 2020a; Zhang
47 et al., 2022). However, Cr (VI) interactions with clay minerals are weak and can be easily disrupted
48 through ion exchange (Covelo et al., 2008; Frank et al., 2019; Rosales-Landeros et al., 2013). Soil
49 organic matter, sulfides, and Fe (II) can convert Cr (VI) to Cr (III), which can then bind to iron oxides
50 or co-precipitate. Soil components like oxides, clay minerals, organic matter, collectively influence the
51 geochemical behavior of Cr in soil systems, with oxides, particularly Fe and Mn oxides, playing a

52 dominant role. The bacterial activity, experimented using soil incubations have shown that weakly
53 crystalline or amorphous Fe oxides can be reduced, leading to the release of associated metals,
54 including Cr (Quantin et al., 2002).

55 Tropical and subtropical climates, characterized by heavy rainfall and high temperatures, result
56 in extensive soil weathering and leaching, causing the formation of "red soils" with phyllosilicates and
57 iron, aluminum, and manganese accumulated as oxides (Berger and Frei, 2014; He et al., 2004). Iron
58 and Mn oxides continually dissolve and precipitate, forming iron-manganese nodules (FMNs), which
59 act as important scavengers of soil pollutants, including Cr. FMNs were found to influence the
60 oxidation of Cr (III) to Cr (VI) (Hai et al., 2020), and field studies indicated that Fe and Mn oxides
61 promoted the migration and transformation of chromium in basalt-derived soils (Ao et al., 2022b; Sun
62 et al., 2022; Wille et al., 2018; Zhang et al., 2023). In naturally Cr-enriched soils, such as those found
63 in the ultramafic areas in Brazil and New Caledonia, Cr primarily associates with well-crystallized Fe
64 oxides or chromite, indicating the predominance of stable trivalent chromium. However, notable
65 amounts of hexavalent chromium (plant-available, up to 1,000 mg/kg) have been detected. This
66 occurrence may be linked to the oxidation of Cr (III) by Mn oxides (Becquer et al., 2003; Fandeur et
67 al., 2009; Garnier et al., 2013, 2006), influenced by the use of fertilizers, as Cr (VI) can exchange with
68 phosphates (Becquer et al., 2003), or associated with microbial weathering (Quantin et al., 2002). In
69 floodplain soil profiles, prolonged flooding increases chromium mobility, while chromium solubility
70 is higher below the water table, highlighting the role of soil hydrology in chromium dissolution and
71 migration (Shaheen and Rinklebe, 2014). A gap remains between controlled lab experiments and field
72 studies with more variables impacting chromium dynamics.

73 Non-traditional isotope techniques have emerged to understand metals and metalloid cycles in
74 modern or paleoenvironments. In red soils, significant isotope fractionation, like that of Cd and Mg
75 isotopes, reveals formation processes linked to FMNs, originating from soil water or granite
76 weathering (Gao et al., 2021, 2018). Redox activity dictates Cr mobility, highlighting the chromium
77 isotope system as a valuable proxy for understanding red soil formation and Cr fate. Recent findings
78 emphasize that redox reactions, particularly the reduction of hexavalent chromium to trivalent
79 chromium, are primary contributors to chromium isotope variability in surface environments (Qin and

80 Wang, 2017; Wang et al., 2021). Studies have evaluated the effect of single or multiple environmental
81 components on chromium redox and isotopic fractionation. Biotic or abiotic reduction of Cr (VI) to
82 Cr (III) results in enrichment in light isotopes of Cr (III) ($\Delta^{53}\text{Cr} = \delta^{53}\text{Cr}_{\text{product}} - \delta^{53}\text{Cr}_{\text{reactant}}$, varying
83 between -7.6‰ and -0.4‰), whereas oxidation of Cr (III) results in a variable $\Delta^{53}\text{Cr}$ (from -2.5‰ to
84 +1‰) (Qin and Wang, 2017). However, the understanding of isotopic fractionation during Cr (III)
85 oxidation is hampered by limited experimental investigations and an incomplete understanding of the
86 underlying mechanisms.

87 Cr isotope records in current soil profiles are limited compared to extensive research on ancient
88 paleosols indicating atmospheric oxygenation (Babechuk et al., 2019; Colwyn et al., 2019; Mänd et
89 al., 2022). Berger and Frei (2014) have characterized Cr behavior in Malagasy Madagascar's laterite
90 soil, revealing increased Cr concentration in the uppermost portions of the soil profile, accompanied
91 by a decrease in $\delta^{53}\text{Cr}$ values, which is interpreted by a partial immobilization (e.g., by adsorption or
92 co-precipitation of Fe oxides) occurring during the upward migration of mobile Cr (III) from deeper
93 parts of the weathering profile. Co-variation of chromium isotopes with silicon isotopes and iron
94 content, along with significant chromium enrichment and silicon loss found in laterite profile by Wille
95 et al. (2018), underscores the impact of Fe oxides on chromium mobility. In laterite profiles, different
96 $\delta^{53}\text{Cr}$ value distribution is linked to parent rock stratification and groundwater levels fluctuations
97 (Paulukat et al., 2015). The migration of weathering fluids in porous layered rocks can lead to uneven
98 chromium loss and fractionation. Additionally, changes in groundwater levels introduce heterogeneity
99 in the Cr isotopic composition and soil organic matter may contributed to chromium fractionation
100 through its ability to reduce isotopically heavy Cr (VI) to Cr (III).

101 Our study investigates chromium distribution in soil profiles from South China, focusing on
102 paddy and non-paddy soils. These regions are paramount to China's food security strategy, especially
103 in rice cultivation. However, persistent food security challenges have emerged in recent years due to
104 toxic metal contaminants in rice (Huang et al., 2019; Ren et al., 2022). These challenges stem from
105 two primary sources: the naturally elevated geological background, as reported in the karst region of
106 Guangxi (Yang et al., 2021), and anthropogenic activities, notably observed in Hunan Province (Wang
107 et al., 2016). Our research aims to utilize a combination of chemical extraction and chromium stable

108 isotope techniques to understand differences in Cr partitioning between FMNs and surrounding soils.
109 We seek to identify potential chromium sources and explore key factors influencing chromium cycling
110 in red soil systems with abundant FMNs, thereby enhancing our understanding of chromium behavior.
111 This study provides a scientific basis for assessing soil chromium pollution and food safety in the
112 region. Additionally, it offers insights into the applicability of chromium isotope techniques in natural
113 soil systems.

114 **2. Material and methods**

115 2.1 Study area and sampling

116 The study was carried out in Wangling Town (109°01'07"E, 23°11'01"N) and Litang Town
117 (109°05'49"E, 23°11'38"N), in Binyang County, Guangxi Zhuang Autonomous Region, in South
118 China. About 8 km separate the two locations. The subtropical monsoon controls the climate here,
119 resulting in an annual precipitation of ~1,491 mm, with a wet season from April to September,
120 accounting for about 78% of the annual precipitation, and a mean annual temperature of 21°C.
121 Carbonated bedrock dominates the region, but red soils are mostly developed from quaternary
122 colluvium deposited over carbonates (Wei et al., 2014). At each site, two soil profiles were sampled:
123 a paddy (rice culture) and a non-paddy soil profile (vegetable culture). The land use has not been
124 modified from at least 40 years. Based on soil texture and color, soil profiles were separated at both
125 sites into several horizons; see details in Table S1. Contrary to the three other profiles, the Litang paddy
126 soil profile is drained at a depth of 40 cm, the outlet water being exported and collected in an artificial
127 cemented reservoir, where rainfall and drained waters are stored and used for irrigation.

128 For each horizon, approximately 50 kg of bulk soil samples were collected, air-dried, and
129 separated into two parts in the laboratory: FMNs (> 1 mm) and surrounding soil (< 1 mm). The
130 separated FMNs were washed gently with Milli-Q water, dried, and then sieved into three size classes:
131 1-2 mm, 2-5 mm, and > 5 mm. The nodule proportions of each horizon were determined by weighing
132 each FMN class. Finally, 5-10 g of each sub-sample (FMNs and surrounding soil) were ground in an
133 agate mortar, passed through 0.15 mm sieves, and stored in sealed bags for further analysis.
134 Additionally, to constrain natural and anthropogenic inputs in soils, three samples of bare calcareous

135 rocks, and seven samples of fertilizers used by local farmers, as well as two samples of irrigation water
136 at the sampling sites of both locations were collected (Table 2). Rocks and fertilizers were ground,
137 dissolved. Water samples (~12 L) were filtered *in situ* through 0.22 μm polyethersulfone membranes
138 (Jinteng, China) and acidified with distilled 16 M HNO_3 .

139 2.2 Physico-chemical properties of soils

140 All chemicals used in this study were at least analytic grade from Merck company and acids were
141 double-distilled for enhanced purity. Soil properties were determined using bulk soil except for XRD
142 analysis. Soil pH was measured potentiometrically in a soil/water (1:2.5) mixture with a PHS-3C pH
143 meter (Leici, China) (Thomas, 2018). The cation exchange capacity (CEC) was determined by the
144 ammonium acetate saturation technique combined with the titration method (Gillman et al., 1983). Wet
145 sieving in conjunction with the hydrometer technique was used to determine soil texture (Gee and
146 Bauder, 2018). Soil bulk density was estimated by weighing and calculating the dry weight of soil per
147 volume unit (Grossman and Reinsch, 2018). Soil organic matter (SOM) content was evaluated using
148 an oil bath- $\text{K}_2\text{Cr}_2\text{O}_7$ titration method (Nelson and Sommers, 2015). The mineralogical composition of
149 FMNs and surrounding soil (< 0.15 mm fraction) was measured by X-ray diffraction (XRD) analysis
150 using a PANalytical diffractometer with Ni-filtered $\text{Cu K}\alpha$ radiation (at 45 kV and 40 mA) scanning
151 from 5° to $75^\circ 2\theta$ with a scan step of 0.01° at RX platform of Université Paris Cité. The minerals were
152 then identified using X'Pert HighScore Plus software (PANalytical, Almelo, The Netherlands) and
153 quantified using the MAUD program (Material Analysis Using Diffraction) based mainly on the
154 Rietveld refinement method (Lutterotti et al., 1999). Element content was measured by inductively
155 coupled plasma mass spectrometry (ICP-QMS 7900, Agilent) at PARI Analytical Platform of IPGP-
156 Université Paris Cité (hereafter referred to as PARI-IPGP) after HCl-HF-HNO_3 digestion. The USGS
157 geological standards BHVO-2, BCR-2, and Nod-P-1 were used for quality control of sample digestion
158 and elemental analysis. In most cases, the measurement deviation does not exceed 10% of the reference
159 value, and the standard deviation is below 5%.

160 2.3 Fe oxide extraction using the citrate-bicarbonate-dithionite (CBD) method

161 Fe oxides were extracted to examine their impact on Cr isotope signatures due to the affinity of
162 Cr towards Fe oxides. A modification of the citrate-bicarbonate-dithionite (CBD) deferrification
163 method from Holmgren (1967) was used to quantify Cr associated with Fe-oxides. One hundred mg
164 of sample was mixed with 40 mL of 0.3 M sodium citrate, 5 mL of 1 M sodium bicarbonate, and 1 g
165 of sodium dithionite, then heated at 80°C for 4 h with manual shaking every 30 min. After centrifugation,
166 the supernatant was removed while the solid residue was kept. The extraction was repeated 3-10 times
167 to ensure that all Fe oxides had been removed. Finally, the residue was rinsed several times with Mill-
168 Q water and dried at 40°C before being weighed, grounded, and stored before elemental and isotope
169 analyses.

170 2.4 Cr purification and isotope analysis

171 The chromium purification was achieved by a 2-step column chemistry with cation exchange
172 resin (AG®50W-X8, 200-400 mesh, Bio-Rad) using the procedure modified from Larsen et al. (2016)
173 in a Class 1000 cleanroom (PARI-IPGP). Since Cr isotope fractionation can occur during the chemical
174 separation through chromatographic columns and the final purification yields rarely exceed 70-80%,
175 we used a double spike technique following the protocol detailed in Sossi et al. (2018).

176 First, samples containing 1 µg Cr were spiked with an appropriate ⁵⁰Cr-⁵⁴Cr double spike solution
177 volume to achieve a Cr_{spike}/Cr_{sample} ratio of ~0.28. The spiked samples were heated in 6 M HCl at 100°C
178 for 24 h to facilitate homogenization and promote the formation of Cr-Cl complexes. Then, cold Milli-
179 Q water (4 L) was added to dilute the HCl concentration to 1 M. Sample was loaded into a pre-cleaned
180 and conditioned column of 1 mL resin, and Cr was collected immediately. Further elution of Cr was
181 carried out in 5 mL of 1 M HCl, and both eluates were evaporated together. The resin was washed with
182 10 mL of 6 M HCl and 5 mL of Mill-Q water. After that, the Cr fractions were exposed to 0.5 M HNO₃
183 and 0.8% H₂O₂ at room temperature for at least 48 h to facilitate the conversion of all Cr species to
184 Cr³⁺. The remaining major elements (e.g., Al) as well as interferents (Ti, V, and Fe) were removed by
185 loading the sample onto the pre-conditioned AG®50W-X8 resin followed by elution with 5 mL of 0.5
186 M HF and 22 mL of 1 M HCl. Finally, Cr was eluted with 10 mL of 2 M HCl, evaporated, and the Cr
187 concentration was adjusted to ~100 µg/L with 2% HNO₃ for isotope measurement.

188 The Cr isotope compositions were measured on a Neptune Plus MC-ICP-MS (ThermoScientific)
189 equipped with a PFA μ Flow nebulizer (100 μ L/min), an Apex HF desolvation introduction system
190 (Elemental Scientific) and eight Faraday collectors that allowed for simultaneous collection of ^{50}Cr ,
191 ^{52}Cr , ^{53}Cr , ^{54}Cr together with ^{49}Ti , ^{51}V and ^{56}Fe , to correct the isobaric interferences of ^{50}Ti , ^{50}V , and
192 ^{54}Fe on ^{50}Cr and ^{54}Cr , respectively. In the medium resolution mode, a 4-8 V sensitivity on ^{52}Cr was
193 obtained for a 100 $\mu\text{g/L}$ Cr solution using Jet sample and H skimmer cones. The purification yield of
194 our method ($\sim 80\%$) is in line with previous studies (Zhu et al., 2018, 2019). The procedural blanks of
195 0.2-11 ng ($n = 7$), are negligible compared to the loaded sample mass (1000 ng). Isotope results are
196 reported as $\delta^{53}\text{Cr}$ and normalized to the certified Cr isotopic standard NIST SRM 979 as Eq. (1):

$$197 \quad \delta^{53}\text{Cr}(\text{‰}) = \left(\frac{(^{53}\text{Cr}/^{52}\text{Cr})_{\text{sample}}}{(^{53}\text{Cr}/^{52}\text{Cr})_{\text{NIST SRM 979}}} - 1 \right) \times 1000 \quad (1)$$

198 Reproducibility and accuracy of isotope measurements were checked using the NIST SRM 3112a
199 and USGS BHVO-2 reference materials, resulting in $\delta^{53}\text{Cr}$ of $-0.09 \pm 0.03\text{‰}$ (2SD, $n = 9$) and $-0.14 \pm$
200 0.05‰ (2SD, $n = 27$), respectively, which agreed with reported values in the literature (Liu et al., 2019;
201 Wu et al., 2020).

202 3. Results

203 3.1 Soil physicochemical properties

204 The key physicochemical properties of the soils are listed in Table 1. Neutral pH values ($7.2 \pm$
205 0.13) are found in most soil horizons except for the non-paddy soil profile in Wangling (pH ~ 4.8) and
206 the topsoil horizon of the two paddy soil profiles in Wangling (pH 6.2) and Litang site (pH 6.5). The
207 difference in soil pH may be attributed the different crops and the recycling of organic material, which
208 tends to acidify soil (Butterly et al., 2013; Wei et al., 2006). The SOM content is higher in the surface
209 horizon and decreases with depth. Overall, paddy soils have more SOM than non-paddy soils because
210 a small portion of above-ground rice stalks and underground roots remains in the field after rice harvest,
211 where intermittent reducing conditions promote their decomposition. The soil density ranges from 0.9
212 to 2.3 g/cm^3 , with the minimum and maximum values observed in the Litang paddy soil profile.
213 Additionally, there is a gradual increase in soil density with depth within this profile, finally reaching
214 approximately twice the value of the bottom horizon in the Wangling paddy soil profile. Soil CEC and

215 clay content increase with depth, ranging from 11.8 to 24.1 cmol/kg and 16 to 40%, respectively.

216 3.2 Distribution and mineralogy of FMNs and surrounding soils

217 In all soil profiles, except with of the Litang paddy soil profile, FMNs content in each size class
218 and across each soil profile exhibits relatively low variability (Table 1). In particular, both non-paddy
219 soil profiles have a dominant FMNs content (61-76% of the total) over the surrounding soil, while the
220 paddy soil profile at Wangling has a lower FMNs content (37-51%). On the other hand, the Litang
221 paddy soil profile shows a gradually increase in the total FMNs content with depth, ranging from ~5%
222 in the top to 45% in the bottom horizon. This increase affects each FMNs size class. In soil profiles,
223 most FMNs are found in the 2-5 mm class (~66%), followed by the > 5 mm class (~23%), and the 1-
224 2 mm class (~10%).

225 Mineral compositions of FMNs are generally similar independently of size or soil horizon (Fig.
226 S1, Table S2); we now discuss these FMNs as a whole compared with the surrounding soils, as depicted
227 in Fig. 1. In both non-paddy and paddy soil profiles of Wangling, the mineralogy remains constant
228 with depth irrespective of fraction (surrounding soils and FMNs, Figs. 1a and 1c, Table S2).
229 Specifically, the surrounding soil at Wangling is primarily composed of ca. 35-40% kaolinite, 30-35%
230 quartz, 10-15% gibbsite, and 10-15% goethite, whereas the FMNs encompass 60-65% goethite, 15-
231 20% gibbsite, and 10-15% kaolinite. However, distinctions in mineralogical composition are
232 noticeable between non-paddy and paddy soil profiles in Litang. In the Litang non-paddy soil profile,
233 from top to bottom, the surrounding soil is affected by a decrease in quartz (55 to 35 %) and microcline
234 (10 % to traces) and a rise in kaolinite (35 to 65%) proportions (Fig. 1b). A similar vertical evolution
235 of the quartz-kaolinite couple is also observed in the surrounding soil of Litang paddy soil profile (Fig.
236 1d). The 15 % of illite found in the upper horizons are progressively replaced by microcline at
237 intermediate depth. The deepest sample is almost exclusively composed of kaolinite and quartz. FMNs
238 in both profiles are primarily constituted of 60-80 % iron oxides (mostly goethite) and 10% quartz
239 (Fig.1b and 1d). Kaolinite represents 10-30 % of the nodules, except in the waterlogged horizon (Brg,
240 45-70 cm) in the Litang paddy soil profile where it is almost absent. Mn-rich phases (e.g., birnessite)
241 have not been observed in either FMNs or surrounding soil samples, primarily due to their low

242 abundance or to the presence of Mn as impurities in other minerals like iron oxides (Wang et al., 2023).

243 3.3 Vertical distribution of elements in soil profiles

244 The concentrations of Cr and other key elements are listed in Table S3 and plotted in Fig. S2.
245 High Mn concentrations (up to 24 g/kg) are only found in the largest FMNs of two horizons of the
246 non-paddy Litang profiles, showing that overall FMNs should be better defined as iron nodules only.
247 Compared to the surrounding soils, FMNs contain higher redox-sensitive trace elements such as V, Cr,
248 and Fe. Conversely, higher concentrations of immobile elements (e.g., Ti) or major and soluble
249 elements (e.g., Mg) are reported in the surrounding soils. In the surrounding soils, Cr concentrations
250 are higher in Wangling (223-247 $\mu\text{g/g}$) than those in Litang (95-180 $\mu\text{g/g}$) and relatively independent
251 of the cultivation mode (paddy vs. non-paddy). Across all soil profiles, Cr content is higher in the
252 smallest FMNs (822-1,451 $\mu\text{g/g}$) than in the largest ones (325-693 $\mu\text{g/g}$).

253 To evaluate the potential accumulation or depletion of elements throughout the soil profile, the
254 mass balance coefficient (τ) was applied (Braun et al., 1993) according to Eq. (2):

$$255 \tau_{X,Nb} = \frac{(C_X/C_{Nb})_{sample}}{(C_X/C_{Nb})_{parent}} - 1 \quad (2)$$

256 where C_X and C_{Nb} are the concentrations of element X and niobium in soil samples or parent materials,
257 respectively. The corresponding values in the deep Quaternary red earth reported by Yang (2006) were
258 used as those in the parent materials, since our soil profiles were developed from quaternary colluvium.
259 Niobium was selected as an immobile element due to its stability during advanced weathering
260 processes (Guinoiseau et al., 2021). As indicated in Fig. 2, the τ values of immobile elements (Al and
261 Ti) remain constant across all soil profiles, suggesting a similar conservative behavior of Nb. The
262 major and soluble element (Mg) shows a slight depletion. The reactive element (Mn) behavior is
263 comparable to that of Mg in the Wangling soil profiles. In contrast, in Litang soil profiles, Mn displays
264 either pronounced depletion (in the paddy soil profile) or enrichment (in the non-paddy soil profile).
265 The redox-active elements (Cr, V, Fe) show strong enrichment, particularly in the non-paddy soil
266 profiles. Significant enrichment of redox-active elements is observed mainly in the Brg horizon (45-
267 70 cm) downwards in the paddy soil profile of Litang, which contrasts with the trend observed in
268 Wangling.

269 To assess the extent of soil weathered erosion throughout the soil profile, we applied the soil-
270 weathered erosion coefficient (ba), using the content of the following oxides, as described by Zhang
271 et al. (2004):

$$272 \quad ba = \frac{(Na_2O + MgO + K_2O + CaO)}{Al_2O_3} \quad (3)$$

273 This coefficient indicates the degree of leaching of soil base ions. Al_2O_3 is more stable and less
274 susceptible to washing away during soil formation than Na_2O , MgO , K_2O , and CaO . Thus, the greater
275 the leaching effect of soil weathering, the lower the ba value. As shown in Table 1, the ba values are
276 generally low (≤ 0.07), except in the Litang paddy profile where ba values are notably higher, peaking
277 at 0.19 at the surface but gradually decreasing with depth to 0.09 at the bottom. This indicates that the
278 Litang paddy profile experiences weaker and more variable soil-weathered leaching compared to the
279 other profiles.

280 3.4 Effect of Fe oxides on element partitioning via CBD extraction

281 For comparison, the CBD extraction was conducted on the two soil profiles from Litang, to
282 determine: 1) the main carrier of Cr both in FMNs and surrounding soils, 2) the respective Cr isotope
283 signatures of the oxides (FeO hereafter, extracted fraction) and silicate phases (ReS hereafter, residual
284 fraction), and 3) whether the observed $\delta^{53}Cr$ values shape in the Litang paddy soil profile is associated
285 with Fe oxide or silicate phases. The paddy soil profile has a heterogeneous elemental distribution
286 (Table S3 and Fig. S2) and the lowest and the most variable FMNs content (Table 1) with depth. In
287 contrast, the non-paddy soil profile is characterized by a homogeneous elemental distribution and a
288 higher and more stable FMNs content with depth. In addition to the surrounding soil fraction, the
289 medium-sized FMNs (2-5 mm) were also included since this fraction represents 58-71% of the total
290 FMNs mass in these profiles. The results are reported in Fig. 3 and Table S4. Specifically, the ReS
291 fraction of the surrounding soil and FMNs has similar low Cr concentrations (60-130 mg/kg, Fig. 3a)
292 in non-paddy and paddy soils. The FeO fraction of the FMNs contains the highest Cr concentrations
293 (800-1000 mg/kg), while Cr concentrations in the FeO fraction of the surrounding soil are intermediate
294 (200-400 mg/kg). Additionally, in the 2-5 mm FMNs, the CBD extraction removed almost all Fe, Mn,
295 and Cr ($> 90\%$), indicating that Cr is almost entirely carried by Fe and Mn oxides in FMNs. In the

296 surrounding soils, however, the extraction yield of Cr (50-75%) is lower than for Fe and Mn (80-95%),
 297 demonstrating that Cr is carried by oxides and silicates (illite, kaolinite or microcline, cf. Fig. 1).
 298 Further, although the Fe extraction yield is almost constant in the Litang paddy soil profile (80-85%),
 299 the Cr and Mn extraction yields appear to be at their maximum in the waterlogged horizon (Brg, 45-
 300 70 cm). This may imply that the waterlogged conditions in this horizon strongly influences the
 301 behavior of Cr and Mn.

302 3.5 Cr isotope signature in soil profiles

303 Chromium isotope signatures of FMNs and soil samples are reported in Fig. 4 and Table S3.
 304 Heavier Cr isotope signatures are observed in FMNs compared with the surrounding soils (Figs. 4a,
 305 4b, 4d and 4e). Between the different FMNs size classes, Cr isotope signature does not vary much,
 306 with a mean value of $0.78 \pm 0.17\text{‰}$ ($n = 45$), which is not correlated with the enrichment of Cr in the
 307 smallest FMNs (Table S3 and Fig. S2). Regarding the surrounding soils, $\delta^{53}\text{Cr}$ values are constant and
 308 cluster around $0.32 \pm 0.07\text{‰}$ across all soil profiles except for the Litang paddy soil profile. Conversely,
 309 the surrounding soils of Litang paddy soil profile demonstrate a clear increase in $\delta^{53}\text{Cr}$ values with
 310 depth, ranging from $0.02 \pm 0.05\text{‰}$ in the topsoil to a maximum of $0.38 \pm 0.06\text{‰}$ in the waterlogged
 311 horizon (Brg horizon, 45-70 cm), and decreasing to $0.19 \pm 0.03\text{‰}$ in the bottom horizon (Fig. 4e and
 312 Table S3).

313 The chromium isotope signatures of the two fractions isolated by CBD extraction differ
 314 significantly (Table S4 and Figs. 4c, 4f). Because of the high Na concentration in the FeO extract, the
 315 analysis of chromium isotope signatures associated with iron oxides cannot be performed directly.
 316 Therefore, to determine the $\delta^{53}\text{Cr}_{\text{FeO}}$, we calculated the extracted mass of iron oxides (Mass_{FeO}) and its
 317 element content (X_{FeO}), assuming that Nb is resistant to CBD extraction and is concentrated in the
 318 residue:

$$319 \quad Nb_{\text{Res}} \times (\text{Mass}_{\text{total}} - \text{Mass}_{\text{FeO}}) = Nb_{\text{total}} \times \text{Mass}_{\text{total}} \quad (4)$$

320 then the extracted ratio and X_{FeO} were calculated according to:

$$321 \quad \text{Mass}_{\text{FeO}} / \text{Mass}_{\text{total}} = 1 - Nb_{\text{total}} / Nb_{\text{Res}} \quad (5)$$

$$322 \quad X_{\text{FeO}} = (X_{\text{total}} - X_{\text{Res}}) \times \text{Mass}_{\text{total}} / \text{Mass}_{\text{FeO}} + X_{\text{Res}} \quad (6)$$

323 and by combining the known Cr concentration and isotope composition in the total material and in the
324 ReS fraction, we determined the $\delta^{53}\text{Cr}_{\text{FeO}}$ according to Eq. (7):

$$325 \quad \delta^{53}\text{Cr}_{\text{FeO}} = \frac{(Nb_{\text{ReS}} \times Cr_{\text{total}} \times \delta^{53}\text{Cr}_{\text{total}} - Nb_{\text{total}} \times Cr_{\text{ReS}} \times \delta^{53}\text{Cr}_{\text{ReS}})}{(Nb_{\text{ReS}} \times Cr_{\text{total}} - Cr_{\text{ReS}} \times Nb_{\text{total}})} \quad (7)$$

326 As shown in Fig. 4c, in the non-paddy soil profile of Litang, the $\delta^{53}\text{Cr}$ values are not only constant
327 with depth but also quite similar for ReS (mean: $0.23 \pm 0.06\text{‰}$), for FeO fractions (mean: $0.33 \pm 0.11\text{‰}$)
328 of the surrounding soils and for ReS fraction extracted from the FMNs (mean: $0.36 \pm 0.10\text{‰}$). However,
329 the FeO fraction of the FMNs has a higher $\delta^{53}\text{Cr}$ signature ($0.76 \pm 0.10\text{‰}$), showing a likely isotope
330 fractionation before or during Cr incorporation within Fe oxides. Litang paddy soil profile, however,
331 presents a quite different picture (Fig. 4f). In this profile, the $\delta^{53}\text{Cr}$ values for the FeO fraction are
332 always heavier than the ReS fraction in the FMNs and the surrounding soil, and the $\delta^{53}\text{Cr}$ values for
333 the ReS fraction in FMNs and surrounding soils are stable with depth but with a heavier signature in
334 the former ($\Delta^{53}\text{Cr}_{\text{ReS_FMN-ReS_Soil}}$ of 0.53‰). While the gap between FeO and ReS' $\delta^{53}\text{Cr}$ is constant
335 within the FMNs and close to the one measured in the Litang non-paddy soil profile ($\Delta^{53}\text{Cr}_{\text{FeO-ReS}}$ of
336 $0.35 \pm 0.18\text{‰}$), significant variations are noticed within the surrounding soil. A maximum value of
337 $\Delta^{53}\text{Cr}_{\text{FeO-ReS}}$ is being observed in the waterlogged horizon (Brg, 45-70 cm), suggesting that Cr
338 redistribution through redox processes and/or its coprecipitation within Fe oxides or FMNs may be the
339 main drivers of the observed $\delta^{53}\text{Cr}$ changes.

340 3.6 Water, rock and fertilizer samples

341 The chromium isotope compositions and the element contents of these samples are listed in Table
342 2. Irrigation waters from Wangling and Litang have low chromium contents of 0.07 and 0.28 $\mu\text{g/L}$,
343 respectively. Calcareous rocks exhibit high calcium contents of up to 40% (consistent with their
344 mineral composition of $> 90\%$ dolomite or calcite) and chromium contents of 4 mg/kg. The N-P-K
345 compound fertilizers typically contain 10-20% phosphorus and potassium, with potassium primarily
346 being in the form of potassium chloride, except for F3. The chromium content of the fertilizer ranges
347 from 16 to 26 mg/kg. At the same time F3 where potassium is predominantly in the form of potassium
348 sulfate derived from the dichlorination of potassium chloride, has a lower chromium content of 9
349 mg/kg, likely due to chromium loss during dechlorination.

350 Wangling and Litang irrigation waters have very different $\delta^{53}\text{Cr}$ values, $1.47 \pm 0.06\text{‰}$ and $0.16 \pm$
351 0.02‰ , respectively (Table 2). The irrigation water in Wangling originates from a natural stream, and
352 therefore its $\delta^{53}\text{Cr}$ value and Cr/Ca ratio are similar to those of calcareous rocks. In contrast, the
353 irrigation water in Litang is stored in an artificial cemented reservoir, which primarily serves to collect
354 rainwater during rainfall events and functions as a drainage pond for paddy fields. Consequently, its
355 $\delta^{53}\text{Cr}$ value is similar to the surrounding soils in paddy fields. Calcareous rock samples have generally
356 high $\delta^{53}\text{Cr}$ values (mean $1.01 \pm 0.58\text{‰}$, $n = 3$), while fertilizer samples have $\delta^{53}\text{Cr}$ values ranging from
357 0.1‰ to 0.5‰ , except for F3 ($1.49 \pm 0.12\text{‰}$). The high $\delta^{53}\text{Cr}$ value of F3 can be attributed to the
358 preferential loss of light chromium isotopes during its dechlorination.

359 **4. Discussion**

360 4.1 Origin of chromium in red soils: a source effect?

361 Soil profiles have consistent chromium isotopic signatures for FMNs and most of surrounding
362 soils, except for the Litang paddy soil profile, which shows variability. To identify if a source effect
363 can explain the variations observed in these red soils, we need to assess input and output fluxes and
364 compare them to the average Cr concentrations.

365 As a first step, the average Cr concentration and soil density of soil profiles can be used to
366 calculate soil Cr stocks in the cultivated soil horizon in 0.3 m^3 box (Fig. 5) according to Eq. (8):

$$367 \text{Soil Cr stock (g)} = \text{Soil density (g/cm}^3\text{)} \times \text{Cr (mg/kg)} \quad (8)$$

368 Comparing Cr stock, external inputs (fertilizer, rainfall, and irrigation water) and outputs (harvesting,
369 drainage), per unit area reveals significant differences. As shown in Fig. 5, Wangling non-paddy and
370 paddy soil profiles hold $259 \text{ g}/0.3 \text{ m}^3$ and $176 \text{ g}/0.3 \text{ m}^3$ of Cr, while Litang has $220 \text{ g}/0.3 \text{ m}^3$ and 43
371 $\text{g}/0.3 \text{ m}^3$, respectively. Before to analyse Cr cycle in a whole, we assess the "active" Cr stock, i.e.
372 reactive to current pedological conditions by evaluating in details the mineralogical composition and
373 the primary Cr carriers. In red soils, surrounding soils must be decoupled from FMNs story. Indeed, a
374 significant difference has been measured for the residual $\delta^{53}\text{Cr}$ signature after CBD extraction in the
375 surrounding soil ($\delta^{53}\text{Cr} = -0.02 \pm 0.10 \text{‰}$) and the FMNs ($0.51 \pm 0.09 \text{‰}$) of the Litang paddy soil
376 profile (Fig. 4f), despite a similar mineralogy (abundant kaolinite and quartz, Fig. 1d) and Cr

377 concentrations (Fig. 3a).

378 The most plausible explanation is that these two components, surrounding soils and FMNs, are at
379 different formation stages, with FMNs serving as ancient archives that persist. This aligns with findings
380 in the literature, where evidences like old ^{14}C formation ages for FMNs (Pai et al., 2003) and insights
381 from 3-D microstructure analysis (Yu et al., 2020b) support this view. Therefore, FMNs should be seen
382 as historical and chemically stable archives unaffected by current pedological conditions, rather than
383 as recent and reactive mineral phases. In surrounding soils, Cr can be associated with primary minerals
384 like illite/microcline or secondary products such as kaolinite and gibbsite in Wangling. Despite no iron
385 oxides has been detected by XRD analysis, amorphous phases like ferrihydrite can constitute a notable
386 portion (10-30 wt%) based on Fe concentrations in bulk samples (3.8-9.4 wt%, Table S3) and the
387 effectiveness of Fe removal by CBD extraction reported for Litang profiles (80-95%, Fig. 3b). The
388 high percentage (50-75 wt%) of chromium removed by CBD extraction in surrounding soils suggests
389 that these amorphous iron oxides are a major carrier of Cr in this fraction (200-400 $\mu\text{gCr/g}$, Fig. 3a
390 and Table S4). The Cr concentration in the ReS fraction (60-80 $\mu\text{gCr/g}$) and the prevalence of quartz
391 and kaolinite indicate the latter as a significant Cr-carrying phase in red soils. A simple mass balance
392 gives a 120-350 $\mu\text{gCr/g}$ kaolinite, larger than the UCC value (92 $\mu\text{g/g}$, Rudnick and Gao (2014)), but
393 in line with the immobile behaviour of Cr during chemical weathering. Thus, if the surrounding soil is
394 considered as the Cr source available for chemical weathering (formation of kaolinite/gibbsite) or
395 oxidative dissolution (iron oxide precipitation), this “active” Cr stock amounts to 22 and 32 $\text{g}/0.3 \text{ m}^3$
396 in the Litang non-paddy and paddy soil profiles, respectively (Fig. 5).

397 Regarding the input, atmospheric deposition, including rainwater and dust, can be a significant
398 source of soil metal contamination (Fu et al., 2021). Rainwater in the area has low Cr levels (≤ 1 ppb)
399 and with an annual rainfall of $\sim 1,491$ mm (Gao et al., 2021), the maximum rainwater Cr flux of 1.5
400 $\text{mg}/\text{m}^2/\text{yr}$ (F_{rain} , Fig. 5). Human-made dust deposition is negligible due to the absence of factories.
401 Natural dust deposition, following Kok et al. (2021), ranges from 1 to 3 $\text{g}/\text{m}^2/\text{yr}$, giving a Cr flux of
402 0.09 to 0.28 $\text{mg}/\text{m}^2/\text{yr}$ (F_{dust} , Fig. 5), about ten times lower than rainwater input.

403 Irrigation water, as a Cr source in soil, differs between Wangling and Litang. In Wangling, river-
404 based irrigation with low Cr content (0.07 $\mu\text{g/L}$) yields a Cr flux (F_{irw} , Fig. 5) of 0.1 $\text{mg}/\text{m}^2/\text{yr}$. In

405 Litang, water reservoirs receive rainwater and drained water, resulting in a higher Cr concentration
406 (0.28 $\mu\text{g/L}$) due to soil interactions, leading to a Cr flux of 0.2 $\text{mg/m}^2/\text{yr}$ for non-paddy fields. However,
407 for the Litang paddy soil profile, this water cannot be considered an external input as most of the Cr
408 originates from drainage within the soil.

409 Fertilizers impact soil Cr levels. We analyzed seven local fertilizers (Table 2) used for rice and
410 vegetable cultivation. On average, 1,500 kg/ha/yr of fertilizer is applied in paddy fields, and 1,500-
411 3,000 kg/ha/yr in non-paddy fields. With fertilizers containing 20 mg/kg Cr, estimated Cr flux is 3-6
412 $\text{mg/m}^2/\text{yr}$ (F_{fer} , Fig. 5). Notably, widespread fertilizer uses in China began after 1980. Over 40 years,
413 it contributes about 0.12-0.24 gCr/m^2 , less than 1.1% of the soil's "active" Cr stock. Fertilizers are not
414 the primary Cr source in the soil.

415 Cr outputs from soil profiles are limited to crop uptake and runoff. Crop uptake comprises
416 vegetables and rice for non-paddy and paddy soil profiles. In Guangxi Zhuang Autonomous Region,
417 vegetable yield averages 2.2 $\text{kg/m}^2/\text{yr}$ (Ministry of Agriculture, 2016), with a Cr concentration of 0.3
418 mg/kg , resulting in a Cr output of about 0.7 $\text{mg/m}^2/\text{yr}$ (F_{veg} , Fig. 5). For Cr export through rice
419 harvesting, most of the Cr remains in rice roots after harvest. With a Cr concentration of 0.2 mg/kg in
420 rice grains (Xiao et al., 2018) and an aboveground plant biomass of 1.2 kg/m^2 for rice (Roy et al.,
421 2012), the estimated Cr export from rice cultivation is approximately 0.5 $\text{mg/m}^2/\text{yr}$ (F_{rice} , Fig. 5).

422 Estimating Cr runoff is challenging. In Litang, due to the absence of major water bodies, it is
423 likely negligible. In Wangling, the river draining the fields shows low Cr levels and isotope features
424 aligned with calcareous formations within the basin, suggesting a negligible Cr flux outside the field
425 area.

426 Input and output fluxes are depicted in Fig. 5. Fertilizers' impact over 40 years is negligible at the
427 soil formation scale. The net Cr enrichment in soils ranges from 1.0 to 1.4 $\text{mg/m}^2/\text{yr}$ based on long-
428 term inputs and outputs. This rate would require 15,700 to 32,000 years to build the "reactive" Cr pool
429 in surrounding soils, inconsistent with its recent nature. Hence, the primary Cr source in soils is likely
430 the chemical weathering of colluvium parent minerals, influenced by changes in redox conditions.

431 4.2 Land use or hydrological control on elemental distribution in soils

432 Distinct land use may play on soil hydrology, which could influence Cr mobility or distribution
433 within paddy and non-paddy soils. Mineralogically, all soil profiles comprise FMNs rich in goethite
434 (60-85%, Fig. 1) and some accessory minerals (quartz, kaolinite, and gibbsite). The FMNs are more
435 abundant in non-paddy soil profiles than in paddy ones, and only the Litang paddy soil profile shows
436 a variability of FMNs content with an increase in depth. The surrounding soils are mostly an
437 assemblage of quartz and kaolinite, but gibbsite and iron oxides can be found at Wangling sites (Figs.
438 1 and S1). At Litang, illite or microcline are found in the top horizons, while amorphous iron oxides,
439 not detected by XRD analysis, can reach 10-30 % (see section 4.1). In modern soils, illite is frequently
440 found under moderate weathering intensity compared with gibbsite and kaolinite, typical of intense
441 chemical weathering, such as faced red soils (Churchman et al., 2012). Gibbsite typically occurs in
442 soils with insufficient Si supplies under intense leaching, while kaolinite forms in soils with weak
443 leaching and high Si availability (Berger and Frei, 2014; Churchman et al., 2012). Our data suggest
444 that soils in Wangling are facing more extreme chemical weathering than in Litang, as confirmed by
445 the lower ba values observed in Wangling. Therefore, the location (Wangling vs. Litang) rather than
446 the land use (vegetable vs. rice culture) seems to dictate the soil hydrology that will condition the
447 mineralogy and the distribution of elements throughout the soil profile. The lower elevation at
448 Wangling (94 m, Table S1) than at Litang (101 m) may result in a shallower groundwater level,
449 subjecting the entire soil profile to long-term fluctuations in groundwater and consequently enhancing
450 soil leaching intensity. In contrast, in Litang, prominent gleying features are observed within the Brg
451 horizon (45-70 cm) of the paddy soil profile, resulting from localized waterlogging. The soil density
452 can be a controlling factor here since its values at the bottom of the Litang paddy soil profile are
453 approximately twice that of Wangling. Consequently, the compactness of the soil at the bottom of the
454 Litang paddy soil profile acts as a barrier, restricting or slowing down the downward movement of
455 water. This restricted permeability contrasts to the porous and permeable nature of the loose soil in the
456 bottom horizon of the Wangling paddy soil profile. Another controlling factor may be the water
457 drainage at 40 cm in the Litang paddy soil profile. Adjacent to this profile lies a cemented water
458 reservoir, and a drainage pipe is connected to the Litang soil profile, positioned approximately 40 cm
459 below the ground, to collect drainage from it. As a result, the upper 40 cm of this profile, including the

460 Ap1, Ap2, and AB horizons, remain consistently dry during periods of non-rice cultivation, in stark
461 contrast to the horizons below, which exhibit gleying features. Therefore, hydrological conditions are
462 a key driver shaping the red soil profiles. By noticing the distinct behaviour of the Litang paddy soil
463 profile, it is also clear that local change in hydrological conditions influences the soil redox conditions
464 and thus plays on the distribution of redox-sensitive elements such as Cr. The principal component
465 analysis (PCA, Fig. S3) based on element concentrations of soil profiles confirms that redox-sensitive
466 elements (Fe, Cr, As, and V) behave similarly, as already observed in other studies (Aide, 2005; Ettler
467 et al., 2017), and consistent with a redox control. Based on the reactivity of Cr isotopes to redox
468 changes (e.g., Wei et al. (2020)), such a redox control should be seen in our soil Cr isotope signatures.

469 4.3 Highlighting the effect of redox cycling on Cr mobility using Cr isotopes (Litang soils)

470 The link between redox conditions and Cr isotope variations has been reported in numerous
471 studies on paleosols and laterites (i.e., red soils). A typical negative $\delta^{53}\text{Cr}$ values is due to the loss of
472 Cr through oxidative weathering (Berger and Frei, 2014; Crowe et al., 2013). This is attributed to the
473 preferential oxidation of the heavy isotopes of Cr during the conversion of Cr (III)-bearing minerals to
474 more soluble and mobile Cr (VI), leading to residual materials enriched in light isotopes. Meanwhile,
475 the positively fractionated mobile pool of Cr (VI) may be transported via runoff to water bodies, as
476 supported by the isotopically heavier $\delta^{53}\text{Cr}$ values of rivers and oceans (D'Arcy et al., 2016; Sun et al.,
477 2019). However, positive $\delta^{53}\text{Cr}$ values and Cr enrichment have also been observed in some weathering
478 profiles, which have been explained as the re-deposition of the isotopically heavy fraction of Cr (VI)
479 (Berger and Frei, 2014; Crowe et al., 2013).

480 Taken as a bulk, our red soil profiles all exhibit more positive $\delta^{53}\text{Cr}$ values (bulk samples all
481 cluster around 0.7‰, except for the mildly positive values (~0.3‰) in the upper part of the Litang
482 paddy soil profile, Fig. 4) than the $\delta^{53}\text{Cr}$ reported for soils from different geobackgrounds (as
483 summarized in Table S5). Remarkably, all FMNs have a constant enrichment in Cr, Fe, and Mn as well
484 as a similar heavy $\delta^{53}\text{Cr}$ signature, but clear correlations between $\delta^{53}\text{Cr}$ values and the inverse of Cr or
485 Fe concentration exist (*Pearson coefficients* $r = -0.91$ and -0.93 , respectively), and are mostly driven
486 by the soil samples from the Litang paddy soil profile (Fig. 6). A robust correlation between Cr and Fe

487 enrichments has been identified in other weathered soil profiles, where Fe oxides (e.g. goethite,
488 hematite) serve as the primary Cr-bearing minerals, exerting control over the mobility and
489 redistribution of chromium (Becquer et al., 2003; Garnier et al., 2013, 2006; Sun et al., 2022). The
490 weaker correlation between $\delta^{53}\text{Cr}$ values and the inverse of Mn concentration (*Pearson coefficients* r
491 $= -0.74$, Fig. S4) can be attributed to the fact that Mn is likely driving the release of Cr, from minerals
492 through the oxidation, inducing a decoupling of Cr and Mn concentrations. Reversely, Fe serves as a
493 key driver for the sequestration of Cr, effectively immobilizing the released Cr within Fe oxides. This
494 statement is validated by the CBD extraction performed on both surrounding soils and FMNs of the
495 Litang soil profiles, where the Cr concentration in the FeO fraction is much higher than that in the ReS
496 (Fig. 3a), indicating that the majority of Cr is carried by Fe oxides. It is particularly valid in FMNs,
497 where almost all Cr is bound or within goethite, while in the surrounding soils, Cr is significantly
498 associated to silicate minerals like kaolinite (Figs. 1 and 3b). The isotope signatures of the two Cr pools
499 (ReS and FeO) indicates a systematic enrichment in heavy isotopes in the FeO fraction for both
500 surrounding soils and FMNs (Figs. 4c and 4f). In the section 4.1, we have concluded that Cr
501 sequestered in Fe oxides is mostly resulting from ancient (in FMNs), or recent (in surrounding soils)
502 parent material weathering. We thus propose that the enrichment in heavy isotopes in iron oxides is
503 directly linked with a pool of heavy Cr (VI) mobilized by oxidative weathering in our red soil profiles,
504 and which is largely coprecipitated, as other researchers have speculated (Berger and Frei, 2014;
505 Crowe et al., 2013). However, the magnitude of the Cr isotope fractionation between FeO and ReS is
506 less variable for FMNs throughout the profile ($\Delta^{53}\text{Cr}_{\text{FeO-ReS}}$ of ca. 0.35‰) than for the surrounding
507 soils ($\Delta^{53}\text{Cr}_{\text{FeO-ReS}}$ ranging from 0.0 to 0.6‰). Based on the distinct history of FMNs and surrounding
508 soils (see section 4.1), it points towards different scenarios of Cr precipitation within Fe oxide phases.

509 Firstly, FMNs are initially formed as pore fillings in soils (Yu et al., 2020b), and consist of Fe
510 from the surrounding soil matrix and the Fe oxides newly formed through the oxidation and
511 precipitation of Fe^{2+} in the soil solution revealed by Fe isotopes (Feng et al., 2018; Sipos et al., 2023).
512 As a result, FMNs may encapsulate Cr from the soil matrix and adsorb/co-precipitate Cr from the
513 solution (mainly as Cr (VI) species like CrO_4^{2-} , HCrO_4^- , $\text{Cr}_2\text{O}_7^{2-}$) during Fe oxide formation. Numerous
514 studies have shown that significant amounts of Cr (VI) can be adsorbed and incorporated into the

515 structure during the formation and transformation of Fe oxides, and the presence of organic matter and
516 Mn (II) can enhance the adsorption (Liang et al., 2021a; Yu et al., 2020a). Adsorption of Cr (VI) by
517 goethite does not cause significant isotope fractionation according to Ellis et al. (2004), thus Cr (VI)
518 is quantitatively sequestered without isotopic fractionation in the FMNs. These findings indicate that
519 the isotopically heavy Cr (VI) from the soil solution is the most likely source of the Fe oxide-associated
520 Cr in these FMNs. This also explains why FMNs of varying sizes have similar Cr isotopic
521 characteristics despite differences in Cr content. We first hypothesized that the larger surface area of
522 smaller FMNs might lead to an increase of chromium absorption compared to larger nodules. However,
523 scanning electron microscopy pictures (unpublished data) revealed an absence of circular or ring-like
524 structures within the FMNs, and the external layers displayed no noticeable chromium enrichment.
525 FMNs probably formed in one step rather than gradually evolving from smaller to larger sizes. This
526 suggests that higher chromium concentrations might have been found in the soil solutions during the
527 genesis of the smaller FMNs, thereby facilitating chromium accumulation. To validate this hypothesis,
528 further experiments such as CBD extraction from FMNs of varying sizes and the application of
529 methodologies like C^{14} are essential. Regarding the FMNs stability with time, Cornu et al. (2009)
530 showed that Fe oxides in FMNs were not dissolved by redox shifts. Additionally, Quantin et al. (2002)
531 demonstrated that well-crystalline Fe oxides resist to bacterial reduction and reducing conditions,
532 while amorphous and poorly crystalline Fe oxides do not. Given that Cr in our FMNs is mostly carried
533 by crystalline Fe oxides (i.e. goethite, Fig. S1), the FMNs may be stable against soil redox oscillation
534 at later stages, thus minimizing the risk of Cr release. However, further dissolution experiments are
535 needed to confirm this hypothesis, as FMNs from different soils or regions display distinct features
536 and may exhibit variations in their dissolution behavior.

537 Secondly, for the surrounding soils, where iron oxides are amorphous or weakly crystalline, the
538 $\Delta^{53}Cr_{FeO-ReS}$ is almost null in the non-paddy soil profile while it is variable and maximal in the
539 waterlogged horizon (Brg) in the paddy soil profile. This limited $\Delta^{53}Cr_{FeO-ReS}$ could potentially be
540 attributed to fractionation caused by preferential uptake of Cr (VI) by surface vegetation. This uptake
541 may occur through phosphate or sulfate carriers due to the structural resemblances they share with Cr
542 (VI) (Ao et al., 2022a), finally leaving behind Cr (III) enriched in lighter isotopes in the soil. However,

543 this hypothesis fails to elucidate the observed variations in $\delta^{53}\text{Cr}$ values within non-cultivated horizons,
544 where the penetration of plant roots is restricted. Moreover, as discussed earlier, we have demonstrated
545 that the Cr output flux caused by plant uptake does not have a significant impact on soil Cr stock and,
546 thus, should have a limited influence on the soil chromium isotope signatures.

547 Alternatively, the presence of more soluble Cr (III) species, such as organic-Cr (III) complexes,
548 in the surrounding soils might also contribute to the limited $\Delta^{53}\text{Cr}_{\text{FeO-ReS}}$. Previous researches have
549 demonstrated that soil acidity and organic matter facilitate the dissolution of Cr (III)-bearing minerals
550 (Li et al., 2022; Sun et al., 2022).

551 For the Litang paddy soil profile, the limited $\Delta^{53}\text{Cr}_{\text{FeO-ReS}}$ of the surrounding soils in the upper
552 part may be attributed to the unique drainage characteristics of this profile, as discussed in the section
553 4.2. This specific drainage system could have led to the removal of a portion of isotopically heavy Cr
554 (VI) from the soil solution in the upper part of the profile through drainage. Consequently, the soil iron
555 oxides in this region are unable to capture this portion of Cr, resulting in lighter $\delta^{53}\text{Cr}$ values and lower
556 Cr concentrations (Fig. 3a) compared to the Brg horizon. Furthermore, Bonneville et al. (2009) and
557 Quantin et al. (2002) established that amorphous and weakly crystalline forms of iron oxides exhibit
558 greater solubility compared to their crystalline counterparts. Soil incubation experiments have
559 demonstrated that during anaerobic phases, the reduction of Fe (III) oxides not only releases
560 structurally bound chromium but also induces the breakdown of soil aggregates and the dissolution of
561 organic matter. This process indirectly mobilizes chromium (Zhang et al., 2023). Conversely, in
562 aerobic conditions, the oxidation of Fe (II) to form Fe (III) oxides, followed by subsequent
563 recrystallization, initiates the reaggregation of soil colloids, leading to the incorporation of chromium.
564 Therefore, the variability of $\delta^{53}\text{Cr}$ within the FeO fraction of the surrounding soils in this profile may
565 also be directly linked with the dynamics of redox conditions. Based on their lower crystallinity (as
566 evidenced by the absence of iron oxide minerals detected by XRD, Fig. 1) than in FMNs (Cornu et al.,
567 2009), the Fe oxides within the surrounding soils are expected to exhibit higher solubility and reactivity,
568 potentially leading to variations in $\delta^{53}\text{Cr}$ values. The highest $\delta^{53}\text{Cr}$ value and CBD-extractability of Fe,
569 Mn, and Cr measured in the waterlogged horizon (Brg, 45-70 cm) imply repetitive cycles of iron oxide
570 dissolution, Cr oxidation by Mn oxides and subsequent sequestration on Fe oxides, resulting in

571 elevated $\delta^{53}\text{Cr}$ values. Indeed, it has become clear that Cr (III) oxidation is constrained by its inability
572 to diffuse and be sorbed onto the surface of Mn oxides (Liang et al., 2021b). The waterlogged
573 conditions in this region may facilitate such diffusion and promote the sorption of Cr (III) onto Mn
574 oxides. Meanwhile, the neutral pH conditions prevalent in this area contribute to a negatively charged
575 surface of Mn oxides, enhancing the adsorption of trivalent Cr (e.g., Cr^{3+} , $[\text{CrOH}]^{2+}$), thus the
576 possibility for Cr (III) to be oxidized. However, in Ferralsols derived from ultramafic rocks in Brazil,
577 it has been shown that the Cr (VI) generated through the oxidation of Mn oxides may subsequently
578 adsorb onto the soil matrix (i.e. Fe oxides) and be chemically extractable by KH_2PO_4 . Therefore, this
579 portion of highly mobile Cr (VI), associated with fine Fe oxide particles, may be transported via post-
580 rainfall runoff, potentially causing harmful environmental effects (Garnier et al., 2013). Such a process
581 cannot be ruled out in our surrounding soils giving the instability of Fe oxides, coupled with the
582 reactivity of Mn oxides. Thus, the toxicity and mobility of Cr (VI) underscore potential environmental
583 threats, particularly concerning groundwater. Therefore, further studies, including KH_2PO_4 extraction
584 of Cr (VI), dissolution studies of FMNs vs. surrounding soils, and particularly those employing
585 synchrotron XAS (X-ray absorption spectroscopy) to examine the speciation of Cr, Mn, and Fe, will
586 provide more precise information regarding valence, or speciation. This will greatly enhance our
587 understanding and interpretation of Cr isotope variations and the cycling of chromium within the soil
588 profile.

589 **5. Conclusions**

590 This study represents the first integration of CBD extraction experiments and chromium isotope
591 techniques to investigate the dynamics, redistribution, and chromium isotopic fractionation within
592 FMNs-rich red soil profiles in southwest China, including non-paddy and paddy soil profiles. The
593 chromium isotope characteristics of soil nodules, a distinctive soil fraction, are also presented for the
594 first time. Our results show that Cr in the red soils is primarily of geogenic origin, deriving from the
595 weathering of the colluvium parent minerals. The effect of land use (non-paddy vs. paddy) is modest
596 on the distribution and/or mobility of Cr within the red soils. FMNs of varying sizes consistently
597 exhibit higher Cr concentrations compared to the surrounding soils, along with more positive $\delta^{53}\text{Cr}$
598 values and are considered as pedological archives not reactive to current physicochemical conditions.

599 Iron oxide extraction technique shows that both FMNs and surrounding soils consistently exhibit
600 heavier signatures in the oxide than in the residual fraction. It suggests the significant capture of
601 isotopically heavy Cr (VI) from soil solution by Fe oxides. Additional variability of $\delta^{53}\text{Cr}$ observed in
602 the amorphous iron oxides recently precipitated in an active paddy soil demonstrates the dominant
603 control of soil hydrology on driving the soil redox conditions and thus Cr other redox-sensitive element
604 dynamics. Contrary to FMNs, which are supposed relatively stable under current physicochemical
605 conditions, the long-term persistence of amorphous iron phases in red soils is more questionable. These
606 findings underscore the potential of chromium isotopes as a tool for tracing chromium dynamics within
607 natural soil systems, particularly those abundant in redox features. However, they also indicate that
608 considerable variations in chromium isotope features resulting from these redox processes might
609 obscure information regarding chromium pollution sources. Nonetheless, conducting further
610 investigations, encompassing dissolution experiments, chromium speciation in soil solutions, analysis
611 of chromium isotope characteristics, and acquiring in-situ data on chromium and iron speciation using
612 synchrotron analysis, will significantly enhance our understanding of soil chromium stability and
613 dynamics. Addressing the absence of chromium content and isotope characteristics in vegetation,
614 especially in crops like rice plants, emerges as a crucial avenue for future investigation. Overcoming
615 this limitation will significantly bolster our ability to provide a more direct and comprehensive
616 response to regional food security concerns related to chromium. To achieve this, forthcoming studies
617 could focus on extensive sampling and analysis of chromium content and isotopic signatures in various
618 plant species, particularly those prevalent in regions with heightened chromium exposure. Additionally,
619 conducting controlled greenhouse experiments or field studies aimed at elucidating chromium uptake,
620 translocation, and accumulation mechanisms within plant tissues will be pivotal. These future
621 endeavors hold the promise of shedding light on the role of vegetation as a potential indicator and
622 reservoir of chromium contamination, thereby contributing significantly to strategies aimed at
623 mitigating environmental and agricultural implications linked to chromium pollution.

624 **Acknowledgements**

625 This work was supported by the French National Program EC2CO (Ecosphère Continentale et
626 Côtière). Parts of this work were supported by IPGP multidisciplinary program PARI, and by Paris–

627 IdF region SESAME Grant no. 12015908. This study contributes to the IdEx Université de Paris ANR-
628 18-IDEX-0001. Mrs Qin work was supported by a China scholarship Council grant that covered her
629 salary in France. We are grateful for Mrs Sophie Nowak help to interpret the XRD data. We thank Pr.
630 F. Moynier for providing the Cr double spike standard. Authors thank three anonymous reviewers for
631 their appreciated comments who improved the quality of this manuscript.

632

633

634 **References**

635 Aide, M., 2005. Elemental composition of soil nodules from two alfisols on an alluvial
636 terrace in Missouri. *Soil Sci* 170, 1022–1033.

637 <https://doi.org/10.1097/01.ss.0000187351.16740.55>

638 Ao, M., Chen, X., Deng, T., Sun, S., Tang, Y., Morel, J.L., Qiu, R., Wang, S., 2022a.
639 Chromium biogeochemical behaviour in soil-plant systems and remediation strategies: A critical
640 review. *J. Hazard. Mater.* 424, 127233. <https://doi.org/10.1016/j.jhazmat.2021.127233>

641 Ao, M., Sun, S., Deng, T., Zhang, F., Liu, T., Tang, Y., Li, J., Wang, S., Qiu, R., 2022b.
642 Natural source of Cr (VI) in soil: The anoxic oxidation of Cr (III) by Mn oxides. *J. Hazard. Mater.*
643 433, 128805. <https://doi.org/10.1016/j.jhazmat.2022.128805>

644 Babechuk, M.G., Weimar, N.E., Kleinhanns, I.C., Eroglu, S., Swanner, E.D., Kenny, G.G.,
645 Kamber, B.S., Schoenberg, R., 2019. Pervasively anoxic surface conditions at the onset of the
646 Great Oxidation Event: New multi-proxy constraints from the Cooper Lake paleosol.
647 *Precambrian Res* 323, 126–163. <https://doi.org/10.1016/j.precamres.2018.12.029>

648 Becquer, T., Quantin, C., Sicot, M., Boudot, J.P., 2003. Chromium availability in ultramafic
649 soils from New Caledonia. *Sci. Total Environ.* 301, 251–261. [https://doi.org/10.1016/S0048-9697\(02\)00298-X](https://doi.org/10.1016/S0048-9697(02)00298-X)

651 Berger, A., Frei, R., 2014. The fate of chromium during tropical weathering: A laterite profile
652 from Central Madagascar. *Geoderma* 213, 521–532.
653 <https://doi.org/10.1016/j.geoderma.2013.09.004>

654 Bonneville, S., Behrends, T., Van Cappellen, P., 2009. Solubility and dissimilatory reduction
655 kinetics of iron (III) oxyhydroxides: A linear free energy relationship. *Geochim. Cosmochim.*
656 *Acta* 73, 5273–5282. <https://doi.org/10.1016/j.gca.2009.06.006>

657 Braun, J.-J., Pagel, M., Herbilln, A., Rosin, C., 1993. Mobilization and redistribution of
658 REEs and thorium in a syenitic lateritic profile: A mass balance study. *Geochim. Cosmochim.*
659 *Acta* 57, 4419–4434. [https://doi.org/10.1016/0016-7037\(93\)90492-F](https://doi.org/10.1016/0016-7037(93)90492-F)

660 Butterly, C.R., Baldock, J.A., Tang, C., 2013. The contribution of crop residues to changes
661 in soil pH under field conditions. *Plant Soil* 366, 185–198. [https://doi.org/10.1007/s11104-012-](https://doi.org/10.1007/s11104-012-1422-1)
662 1422-1

663 Churchman, G.J., Lowe, D.J., Zealand, N., 2012. Alteration, formation, and occurrence of
664 minerals in soils. *Handbook of Soil Sciences*. 2nd edition. <https://hdl.handle.net/10289/9024>

665 Colwyn, D.A., Sheldon, N.D., Maynard, J.B., Gaines, R., Hofmann, A., Wang, X., Gueguen,
666 B., Asael, D., Reinhard, C.T., Planavsky, N.J., 2019. A paleosol record of the evolution of Cr
667 redox cycling and evidence for an increase in atmospheric oxygen during the Neoproterozoic.
668 *Geobiology* 17, 579–593. <https://doi.org/10.1111/gbi.12360>

669 Cornu, S., Cattle, J.A., Samouëlian, A., Laveuf, C., Guilherme, L.R.G., Albéric, P., 2009.
670 Impact of Redox Cycles on Manganese, Iron, Cobalt, and Lead in Nodules. *Soil Sci Soc Am J* 73,
671 1231–1241. <https://doi.org/10.2136/sssaj2008.0024>

672 Covelo, E.F., Vega, F.A., Andrade, M.L., 2008. Sorption and desorption of Cd, Cr, Cu, Ni,
673 Pb and Zn by a Fibric Histosol and its organo-mineral fraction. *J. Hazard. Mater.* 159, 342–347.
674 <https://doi.org/10.1016/j.jhazmat.2008.02.025>

675 Crowe, S.A., Døssing, L.N., Beukes, N.J., Bau, M., Kruger, S.J., Frei, R., Canfield, D.E.,
676 2013. Atmospheric oxygenation three billion years ago. *Nature* 501, 535–538.
677 <https://doi.org/10.1038/nature12426>

678 D'Arcy, J., Babechuk, M.G., Døssing, L.N., Gaucher, C., Frei, R., 2016. Processes
679 controlling the chromium isotopic composition of river water: Constraints from basaltic river
680 catchments. *Geochim. Cosmochim. Acta* 186, 296–315.
681 <https://doi.org/10.1016/j.gca.2016.04.027>

682 Dimos, V., Haralambous, K.J., Malamis, S., 2012. A Review on the Recent Studies for
683 Chromium Species Adsorption on Raw and Modified Natural Minerals. *Crit Rev Environ Sci*
684 *Technol* 42, 1977–2016. <https://doi.org/10.1080/10643389.2011.574102>

685 Ellis, A.S., Johnson, T.M., Bullen, T.D., 2004. Using Chromium Stable Isotope Ratios To
686 Quantify Cr (VI) Reduction: Lack of Sorption Effects. *Environ. Sci. Technol.* 38, 3604–3607.
687 <https://doi.org/10.1021/es0352294>

688 Ettler, V., Chren, M., Mihaljevič, M., Drahot, P., Křibek, B., Veselovský, F., Sracek, O.,
689 Vaněk, A., Penížek, V., Komárek, M., Mapani, B., Kamona, F., 2017. Characterization of Fe-Mn
690 concentric nodules from Luvisol irrigated by mine water in a semi-arid agricultural area.
691 *Geoderma* 299, 32–42. <https://doi.org/10.1016/j.geoderma.2017.03.022>

692 Fandeur, D., Juillot, F., Morin, G., Olivi, L., Cognigni, A., Webb, S.M., Ambrosi, J.-P.,
693 Fritsch, E., Guyot, F., Brown, Jr., G.E., 2009. XANES Evidence for Oxidation of Cr (III) to Cr
694 (VI) by Mn-Oxides in a Lateritic Regolith Developed on Serpentinized Ultramafic Rocks of New
695 Caledonia. *Environ. Sci. Technol.* 43, 7384–7390. <https://doi.org/10.1021/es900498r>

696 Feng, J., Pei, L., Zhu, X., Ju, J., Gao, S., 2018. Absolute accumulation and isotope
697 fractionation of Si and Fe during dolomite weathering and terra rossa formation. *Chem. Geol.*
698 496, 43–56. <https://doi.org/10.1016/j.chemgeo.2018.08.018>

699 Frank, A.B., Klaebe, R.M., Frei, R., 2019. Fractionation Behavior of Chromium Isotopes
700 during the Sorption of Cr (VI) on Kaolin and its Implications for Using Black Shales as a
701 Paleoredox Archive. *Geochemistry, Geophys. Geosystems* 20, 2290–2302.
702 <https://doi.org/10.1029/2019GC008284>

703 Fu, Y., Li, F., Guo, S., Zhao, M., 2021. Cadmium concentration and its typical input and
704 output fluxes in agricultural soil downstream of a heavy metal sewage irrigation area. *J. Hazard.*
705 *Mater.* 412, 125203. <https://doi.org/10.1016/j.jhazmat.2021.125203>

706 Gao, T., Ke, S., Wang, S.J., Li, F., Liu, C., Lei, J., Liao, C., Wu, F., 2018. Contrasting Mg
707 isotopic compositions between Fe-Mn nodules and surrounding soils: Accumulation of light Mg
708 isotopes by Mg-depleted clay minerals and Fe oxides. *Geochim. Cosmochim. Acta* 237, 205–222.
709 <https://doi.org/10.1016/j.gca.2018.06.028>

710 Gao, T., Liu, Yuhui, Xia, Y., Zhu, J.M., Wang, Z., Qi, M., Liu, Yizhang, Ning, Z., Wu, Q.,
711 Xu, W., Liu, C., 2021. Cadmium isotope compositions of Fe-Mn nodules and surrounding soils:
712 Implications for tracing Cd sources. *Fundam. Res.* 1, 269–276.
713 <https://doi.org/10.1016/j.fmre.2021.04.002>

714 Garnier, J., Quantin, C., Guimarães, E.M., Vantelon, D., Montargès-Pelletier, E., Becquer,
715 T., 2013. Cr (VI) genesis and dynamics in Ferralsols developed from ultramafic rocks: The case
716 of Niquelândia, Brazil. *Geoderma* 193–194, 256–264.
717 <https://doi.org/10.1016/j.geoderma.2012.08.031>

718 Garnier, J., Quantin, C., Martins, E.S., Becquer, T., 2006. Solid speciation and availability
719 of chromium in ultramafic soils from Niquelândia, Brazil. *J Geochem Explor* 88, 206–209.
720 <https://doi.org/10.1016/j.gexplo.2005.08.040>

721 Gee, G.W., Bauder, J.W., 2018. Particle-size Analysis. pp. 383–411.
722 <https://doi.org/10.2136/sssabookser5.1.2ed.c15>

723 Gillman, G.P., Bruce, R.C., Davey, B.G., Kimble, J.M., Searle, P.L., Skjemstad, J.O., 1983.
724 A comparison of methods used for determination of cation exchange capacity. *Commun Soil Sci*
725 *Plant Anal* 14, 1005–1014. <https://doi.org/10.1080/00103628309367428>

726 Grossman, R.B., Reinsch, T.G., 2018. 2.1 Bulk Density and Linear Extensibility. pp. 201–
727 228. <https://doi.org/10.2136/sssabookser5.4.c9>

728 Guinoiseau, D., Fekiacova, Z., Allard, T., Druhan, J.L., Balan, E., Bouchez, J., 2021.
729 Tropical Weathering History Recorded in the Silicon Isotopes of Lateritic Weathering Profiles.
730 *Geophys. Res. Lett.* 48. <https://doi.org/10.1029/2021GL092957>

731 Hai, J., Liu, L., Tan, W., Hao, R., Qiu, G., 2020. Catalytic oxidation and adsorption of Cr
732 (III) on iron-manganese nodules under oxic conditions. *J. Hazard. Mater.* 390, 122166.
733 <https://doi.org/10.1016/j.jhazmat.2020.122166>

734 He, Z., Zhang, M., Wilson, M.J., 2004. Distribution and Classification of Red Soils in China,
735 in: *The Red Soils of China*. Springer Netherlands, Dordrecht, pp. 29–33.
736 https://doi.org/10.1007/978-1-4020-2138-1_3

737 Holmgren, G.G.S., 1967. A Rapid Citrate-Dithionite Extractable Iron Procedure. *Soil Sci*

738 Soc Am J 31, 210–211. <https://doi.org/10.2136/sssaj1967.03615995003100020020x>

739 Huang, Y., Wang, L., Wang, W., Li, T., He, Z., Yang, X., 2019. Current status of agricultural
740 soil pollution by heavy metals in China: A meta-analysis. *Sci. Total Environ.* 651, 3034–3042.
741 <https://doi.org/10.1016/j.scitotenv.2018.10.185>

742 Johnston, C.P., Chrysochoou, M., 2014. Mechanisms of chromate adsorption on hematite.
743 *Geochim. Cosmochim. Acta* 138, 146–157. <https://doi.org/10.1016/j.gca.2014.04.030>

744 Kok, J.F., Adebisi, A.A., Albani, S., Balkanski, Y., Checa-Garcia, R., Chin, M., Colarco, P.R.,
745 Hamilton, D.S., Huang, Y., Ito, A., Klose, M., Leung, D.M., Li, L., Mahowald, N.M., Miller, R.L.,
746 Obiso, V., Pérez García-Pando, C., Rocha-Lima, A., Wan, J.S., Whicker, C.A., 2021. Improved
747 representation of the global dust cycle using observational constraints on dust properties and
748 abundance. *Atmospheric Chem. Phys.* 21, 8127–8167. <https://doi.org/10.5194/acp-21-8127-2021>

749 Larsen, K.K., Wielandt, D., Schiller, M., Bizzarro, M., 2016. Chromatographic speciation of
750 Cr (III)-species, inter-species equilibrium isotope fractionation and improved chemical
751 purification strategies for high-precision isotope analysis. *J. Chromatogr. A* 1443, 162–174.
752 <https://doi.org/10.1016/j.chroma.2016.03.040>

753 Li, D., Li, G., He, Y., Zhao, Y., Miao, Q., Zhang, H., Yuan, Y., Zhang, D., 2022. Key Cr
754 species controlling Cr stability in contaminated soils before and chemical stabilization at a
755 remediation engineering site. *J. Hazard. Mater.* 424, 127532.
756 <https://doi.org/10.1016/j.jhazmat.2021.127532>

757 Liang, C., Fu, F., Tang, B., 2021a. Mn-incorporated ferrihydrite for Cr (VI) immobilization:
758 Adsorption behavior and the fate of Cr (VI) during aging. *J. Hazard. Mater.* 417, 126073.
759 <https://doi.org/10.1016/j.jhazmat.2021.126073>

760 Liang, J., Huang, X., Yan, J., Li, Y., Zhao, Z., Liu, Y., Ye, J., Wei, Y., 2021b. A review of the
761 formation of Cr (VI) via Cr (III) oxidation in soils and groundwater. *Sci. Total Environ.* 774,
762 145762. <https://doi.org/10.1016/j.scitotenv.2021.145762>

763 Liu, C., Xu, L., Liu, C., Liu, J., Qin, L., Zhang, Z., Liu, S., Li, S., 2019. High-Precision
764 Measurement of Stable Cr Isotopes in Geological Reference Materials by a Double-Spike TIMS
765 Method. *Geostand Geoanal Res* 43, 647–661. <https://doi.org/10.1111/ggr.12283>

766 Lutterotti, L., Matthies, S., Wenk, Hjin., others, 1999. MAUD: a friendly Java program for
767 material analysis using diffraction. *IUCR Newsl. CPD* 21, 14–15.

768 Mänd, K., Planavsky, N.J., Porter, S.M., Robbins, L.J., Wang, C., Kreitsmann, T., Paiste, K.,
769 Paiste, P., Romashkin, A.E., Deines, Y.E., Kirsimäe, K., Lepland, A., Konhauser, K.O., 2022.
770 Chromium evidence for protracted oxygenation during the Paleoproterozoic. *Earth Planet. Sci.*
771 *Lett.* 584, 117501. <https://doi.org/10.1016/j.epsl.2022.117501>

772 Nelson, D.W., Sommers, L.E., 2015. Total Carbon, Organic Carbon, and Organic Matter. pp.
773 539–579. <https://doi.org/10.2134/agronmonogr9.2.2ed.c29>

774 Pai, C.W., Wang, M.K., Chiang, H.C., King, H.B., Hwong, J.-L., Hu, H.T., 2003. Formation
775 of iron nodules in a Hapludult of central Taiwan. *Can J Soil Sci* 83, 167–172.
776 <https://doi.org/10.4141/S02-068>

777 Paulukat, C., Døssing, L.N., Mondal, S.K., Voegelin, A.R., Frei, R., 2015. Oxidative release
778 of chromium from Archean ultramafic rocks, its transport and environmental impact - A Cr
779 isotope perspective on the Sukinda valley ore district (Orissa, India). *Appl. Geochemistry* 59,
780 125–138. <https://doi.org/10.1016/j.apgeochem.2015.04.016>

781 Qin, L., Wang, X., 2017. Chromium isotope geochemistry, in: *Non-Traditional Stable*
782 *Isotopes. Rev Mineral Geochem* 82, 379–414. <https://doi.org/10.2138/rmg.2017.82.10>

783 Quantin, C., Becquer, T., Rouiller, J.H., Berthelin, J., 2002. Redistribution of Metals in a
784 New Caledonia Ferralsol After Microbial Weathering. *Soil Sci Soc Am J* 66, 1797–1804.
785 <https://doi.org/10.2136/sssaj2002.1797>

786 Rauch, J.N., Pacyna, J.M., 2009. Earth's global Ag, Al, Cr, Cu, Fe, Ni, Pb, and Zn cycles.
787 *Global Biogeochem Cycles* 23, 1–16. <https://doi.org/10.1029/2008GB003376>

788 Ren, S., Song, C., Ye, S., Cheng, C., Gao, P., 2022. The spatiotemporal variation in heavy
789 metals in China's farmland soil over the past 20 years: A meta-analysis. *Sci. Total Environ.* 806,
790 150322. <https://doi.org/10.1016/j.scitotenv.2021.150322>

791 Rosales-Landeros, C., Barrera-Díaz, C.E., Bilyeu, B., Guerrero, V.V., Núñez, F.U., 2013. A
792 Review on Cr(VI) Adsorption Using Inorganic Materials. *Am. J. Anal. Chem.* 04, 8–16.
793 <https://doi.org/10.4236/ajac.2013.47A002>

794 Roy, K.S., Bhattacharyya, P., Neogi, S., Rao, K.S., Adhya, T.K., 2012. Combined effect of
795 elevated CO₂ and temperature on dry matter production, net assimilation rate, C and N allocations
796 in tropical rice (*Oryza sativa* L.). *Field Crops Res.* 139, 71–79.
797 <https://doi.org/10.1016/j.fcr.2012.10.011>

798 Rudnick, R.L., Gao, S., 2014. Composition of the Continental Crust, in: *Treatise on*
799 *Geochemistry*, Second Ed., pp. 1–51. <https://doi.org/10.1016/B978-0-08-095975-7.00301-6>

800 Shaheen, S.M., Rinklebe, J., 2014. Geochemical fractions of chromium, copper, and zinc
801 and their vertical distribution in floodplain soil profiles along the Central Elbe River, Germany.
802 *Geoderma* 228–229, 142–159. <https://doi.org/10.1016/j.geoderma.2013.10.012>

803 Shahid, M., Shamsad, S., Rafiq, M., Khalid, S., Bibi, I., Niazi, N.K., Dumat, C., Rashid,
804 M.I., 2017. Chromium speciation, bioavailability, uptake, toxicity and detoxification in soil-plant
805 system: A review. *Chemosphere* 178, 513–533.
806 <https://doi.org/10.1016/j.chemosphere.2017.03.074>

807 Sipos, P., Kovács, I., Barna, G., Tóth, A., Makó, A., Palcsu, L., Kiss, G., Horváth, A., Puskás-
808 Preszner, A., 2023. Iron isotope fractionation during the formation of ferromanganese nodules
809 under different conditions of hydromorphism. *Geoderma* 430, 116286.
810 <https://doi.org/10.1016/j.geoderma.2022.116286>

811 Sossi, P.A., Moynier, F., van Zuilen, K., 2018. Volatile loss following cooling and accretion
812 of the Moon revealed by chromium isotopes. *Proc. Natl. Acad. Sci. U.S.A.* 115, 10920–10925.
813 <https://doi.org/10.1073/pnas.1809060115>

814 Sun, S.S., Ao, M., Geng, K.R., Chen, J.Q., Deng, T.H.B., Li, J.J., Guan, Z.T., Mo, B.L., Liu,
815 T., Yang, W.J., Tang, Y.T., Qiu, R.L., 2022. Enrichment and speciation of chromium during basalt
816 weathering: Insights from variably weathered profiles in the Leizhou Peninsula, South China. *Sci.*
817 *Total Environ.* 822, 153304. <https://doi.org/10.1016/j.scitotenv.2022.153304>

818 Sun, Z., Wang, X., Planavsky, N., 2019. Cr isotope systematics in the Connecticut River
819 estuary. *Chem. Geol.* 506, 29–39. <https://doi.org/10.1016/j.chemgeo.2018.12.034>

820 Thomas, G.W., 2018. Soil pH and Soil Acidity. pp. 475–490.
821 <https://doi.org/10.2136/sssabookser5.3.c16>

822 Wang, L., Jin, Y., Weiss, D.J., Schleicher, N.J., Wilcke, W., Wu, L., Guo, Q., Chen, J.,
823 O'Connor, D., Hou, D., 2021. Possible application of stable isotope compositions for the
824 identification of metal sources in soil. *J. Hazard. Mater.* 407, 124812.
825 <https://doi.org/10.1016/j.jhazmat.2020.124812>

826 Wang, M., Chen, W., Peng, C., 2016. Risk assessment of Cd polluted paddy soils in the
827 industrial and township areas in Hunan, Southern China. *Chemosphere* 144, 346–351.
828 <https://doi.org/10.1016/j.chemosphere.2015.09.001>

829 Wang, Z., Wen, Y., Gou, W., Ji, J., Li, W., 2023. Zn isotope signatures in soil Fe Mn nodules
830 with karst high geochemical background. *Sci. Total Environ.* 882, 163365.
831 <https://doi.org/10.1016/j.scitotenv.2023.163365>

832 Wei, W., Kläbe, R., Ling, H.-F., Huang, F., Frei, R., 2020. Biogeochemical cycle of
833 chromium isotopes at the modern Earth's surface and its applications as a paleo-environment
834 proxy. *Chem. Geol.* 541, 119570. <https://doi.org/10.1016/j.chemgeo.2020.119570>

835 Wei, X., Hao, M., Shao, M., Gale, W.J., 2006. Changes in soil properties and the availability
836 of soil micronutrients after 18 years of cropping and fertilization. *Soil Tillage Res* 91, 120–130.
837 <https://doi.org/10.1016/j.still.2005.11.009>

838 Wei, X., Ji, H., Wang, S., Chu, H., Song, C., 2014. The formation of representative lateritic
839 weathering covers in south-central Guangxi (southern China). *Catena (Amst)* 118, 55–72.
840 <https://doi.org/10.1016/j.catena.2014.01.019>

841 Wille, M., Babechuk, M.G., Kleinhanns, I.C., Stegmaier, J., Suhr, N., Widdowson, M.,
842 Kamber, B.S., Schoenberg, R., 2018. Silicon and chromium stable isotopic systematics during
843 basalt weathering and lateritisation: A comparison of variably weathered basalt profiles in the
844 Deccan Traps, India. *Geoderma* 314, 190–204. <https://doi.org/10.1016/j.geoderma.2017.10.051>

845 Wu, G., Zhu, J.-M., Wang, X., Johnson, T.M., Han, G., 2020. High-Sensitivity Measurement
846 of Cr Isotopes by Double Spike MC-ICP-MS at the 10 ng Level. *Anal. Chem.* 92, 1463–1469.
847 <https://doi.org/10.1021/acs.analchem.9b04704>

848 Xiao, G., Hu, Y., Li, N., Yang, D., 2018. Spatial autocorrelation analysis of monitoring data
849 of heavy metals in rice in China. *Food Control* 89, 32–37.

850 <https://doi.org/10.1016/j.foodcont.2018.01.032>

851 Yang, L., 2006. The physical and chemic characters of ferromanganese nodules in the
852 quaternary red earth in southern China (Master thesis). Zhejiang Normal University. (in Chinese)

853 Yang, Q., Yang, Z., Zhang, Q., Liu, X., Zhuo, X., Wu, T., Wang, L., Wei, X., Ji, J., 2021.
854 Ecological risk assessment of Cd and other heavy metals in soil-rice system in the karst areas
855 with high geochemical background of Guangxi, China. *Sci. China Earth Sci.* 64, 1126–1139.
856 <https://doi.org/10.1007/s11430-020-9763-0>

857 Yu, G., Fu, F., Ye, C., Tang, B., 2020. Behaviors and fate of adsorbed Cr (VI) during Fe (II)
858 -induced transformation of ferrihydrite-humic acid co-precipitates. *J. Hazard. Mater.* 392, 122272.
859 <https://doi.org/10.1016/j.jhazmat.2020.122272>

860 Yu, X., Wang, Y., Zhou, G., Peng, G., Brookes, P.C., Lu, S., 2020. Paleoclimatic fingerprints
861 of ferromanganese nodules in subtropical Chinese soils identified by synchrotron radiation-based
862 microprobes. *Chem. Geol.* 531, 119357. <https://doi.org/10.1016/j.chemgeo.2019.119357>

863 Zhang, K., Yang, Y., Chi, W., Chen, G., Du, Y., Hu, S., Li, F., Liu, T., 2023. Chromium
864 transformation driven by iron redox cycling in basalt-derived paddy soil with high geological
865 background values. *J Environ Sci (China)* 125, 470–479.
866 <https://doi.org/10.1016/j.jes.2021.12.013>

867 Zhang, L., Fu, F., Yu, G., Sun, G., Tang, B., 2022. Fate of Cr (VI) during aging of
868 ferrihydrite-humic acid co-precipitates: Comparative studies of structurally incorporated Al (III)
869 and Mn (II). *Sci. Total Environ.* 807, 151073. <https://doi.org/10.1016/j.scitotenv.2021.151073>

870 Zhang, M., Wilson, M.J., He, Z., 2004. Mineralogy of Red Soils in Southern China in
871 Relation to Their Development and Charge Characteristics, in: *The Red Soils of China*. Springer
872 Netherlands, Dordrecht, pp. 35–61. https://doi.org/10.1007/978-1-4020-2138-1_4

873 Zhu, J., Wu, G., Wang, X., Han, G., Zhang, L., 2018. An improved method of Cr purification
874 for high precision measurement of Cr isotopes by double spike MC-ICP-MS. *J Anal At Spectrom*
875 33, 809–821. <https://doi.org/10.1039/C8JA00033F>

876 Zhu, K., Sossi, P.A., Siebert, J., Moynier, F., 2019. Tracking the volatile and magmatic
877 history of Vesta from chromium stable isotope variations in eucrite and diogenite meteorites.

878

Geochim. Cosmochim. Acta 266, 598–610. <https://doi.org/10.1016/j.gca.2019.07.043>

879

1 Table 1
2 Soil key physicochemical properties.

Profile	Horizon	pH	SOM	Density	CEC	Clay	Slit	Sand	FMNs distribution				ba*
									> 5 mm	2-5 mm	1-2 mm	Total	
			g/kg	g/cm ³	cmol/kg	%			g/kg bulk				
Wangling Non-paddy	A	4.9	30.9	1.5	12.4	37.5	23.2	39.3	177	500	82	759	0.02
	B	4.7	19.8	1.7	17.8	40.0	22.5	37.5	87	613	55	755	0.02
Wangling Paddy	Ap1	6.2	49.7	1.3	15.0	34.0	32.4	33.6	81	262	61	404	0.05
	Ap2	7.3	37.9	1.6	14.9	23.0	39.9	37.1	108	310	45	463	0.05
	AB	7.3	16.9	1.7	11.8	23.0	31.4	45.6	124	334	47	506	0.04
	Br1	7.4	12.0	1.5	21.1	33.0	27.9	39.1	59	264	53	376	0.04
Litang Non-paddy	Br2	7.2	8.4	1.2	20.1	37.0	27.4	35.6	42	231	101	374	0.04
	A	7.0	25.7	1.4	12.0	16.0	33.9	50.1	124	411	75	610	0.07
	AB	7.1	26.5	2.0	15.5	39.0	27.2	33.8	274	421	26	721	0.05
Litang Paddy	B	7.4	15.4	1.7	16.5	36.0	27.5	36.5	254	403	25	681	0.06
	Ap1	6.5	50.3	0.9	15.3	28.0	51.0	21.0	15	36	3	55	0.19
	Ap2	7.3	31.1	1.5	15.9	25.0	51.9	23.1	11	28	5	44	0.19
	AB	7.4	16.6	1.8	20.5	35.0	41.7	23.3	10	33	4	47	0.16
	Brg	7.1	12.7	1.7	23.4	39.0	38.4	22.6	35	107	17	158	0.12
	Br	7.2	6.7	2.3	24.1	39.0	24.4	36.6	124	291	36	451	0.09

3 *ba: the soil-weathered eluviation coefficient. FMNs: iron-manganese nodules.

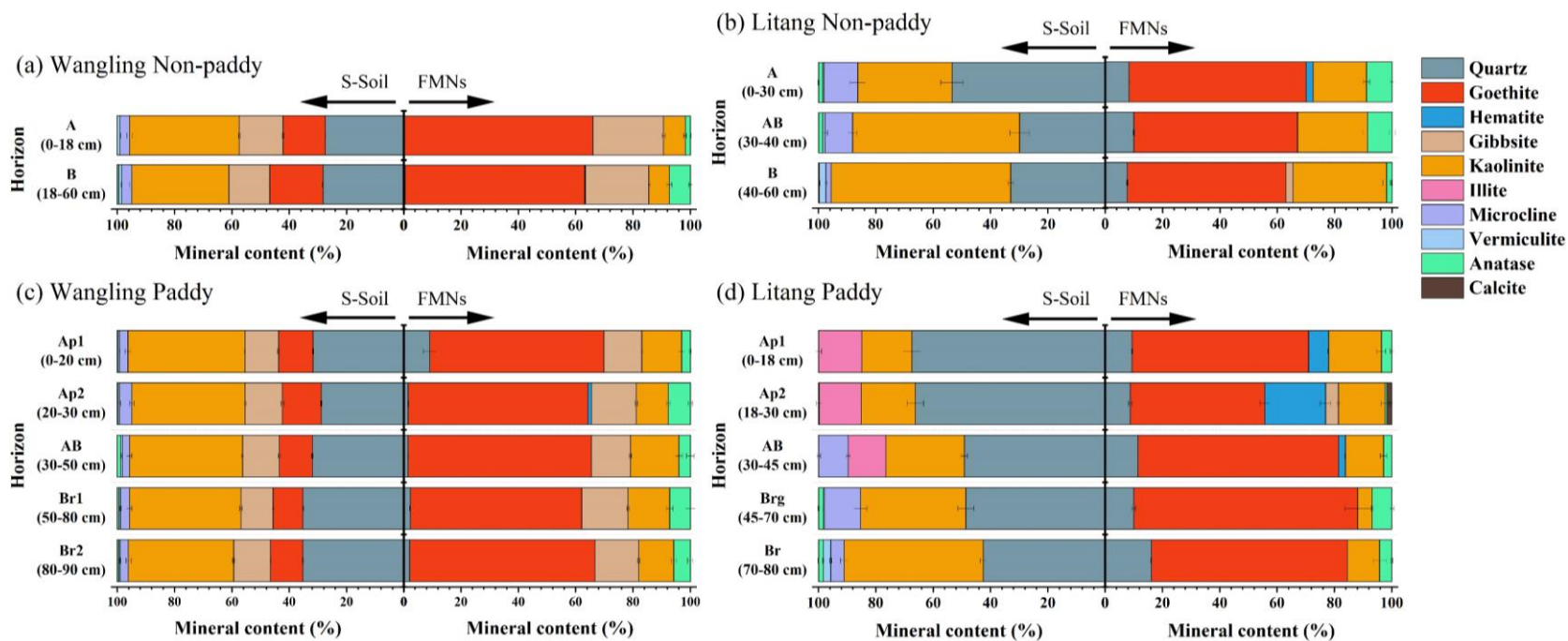
4

5 Table 2

6 Element and Cr isotope compositions of irrigation waters, rocks, and fertilizers.

Element	Irrigation water		Rock			Fertilizer							
	Wangling	Litang	Carbonate1	Carbonate2	Calcite	F0 (NPK 45%, 15-15-15)	F1 (NPK 51%, 17-17-17)	F2 (NPK 45%, 15-15-15)	F3 (NPK 45%, 15-15-15)	F4 (NPK 45%, 15-15-15)	F5 (NPK 45%, 15-15-15)	F6 (NPK 30%, 18-5-7)	
$\delta^{53}\text{Cr} \pm 2\text{SD}$	‰	1.47 ± 0.06	0.16 ± 0.02	1.30 ± 0.07	0.72 ± 0.04	1.01 ± 0.03	0.34 ± 0.01	0.33 ± 0.07	0.36 ± 0.04	1.49 ± 0.12	0.44 ± 0.04	0.11 ± 0.03	0.45 ± 0.04
P	g/kg	bdl	0.001	bdl	bdl	0.002	66.9	72.4	62.5	75.5	62.6	70.1	18.2
K		0.004	0.02	0.003	0.01	0.01	127.7	139.0	133.0	118.9	128.0	117.2	55.9
Ca		0.06	0.06	395.9	395.3	209.5	7.6	8.7	8.8	1.8	9.3	11.4	8.9
Mn		bdl	0.001	2.7	2.7	16.2	724.1	400.7	574.2	88.1	421.0	162.4	100.6
Fe		bdl	0.002	46.8	45.2	175.6	7026.5	4149.2	5323.7	796.8	4353.3	2121.4	2238.3
Cr	µg/kg (ug/L for water samples)	0.07	0.28	3876.8	3836.2	4037.6	19980.5	22300.5	21552.7	8862.1	23074.1	16408.4	25792.8

7 bdl=blow detection limit.



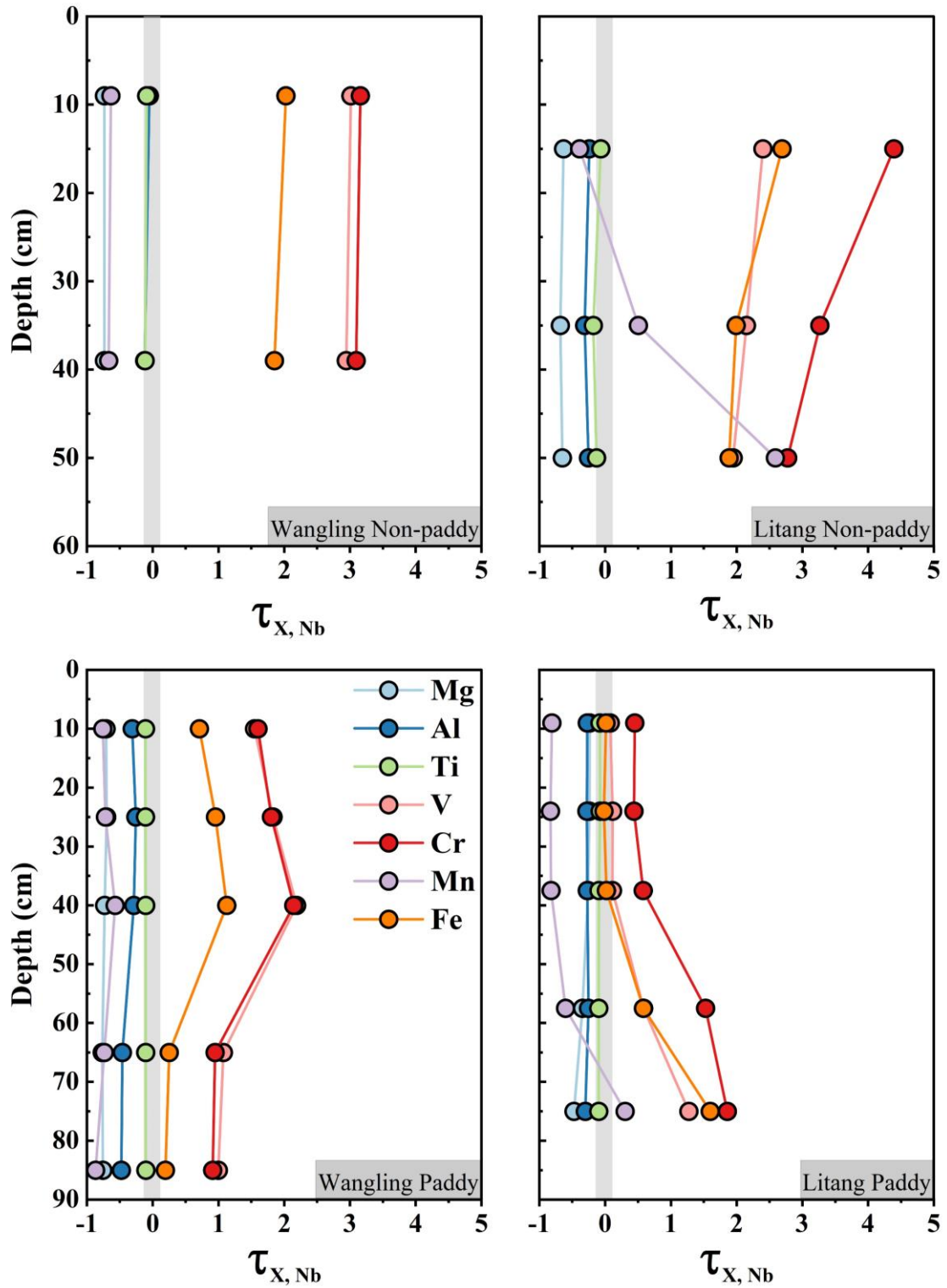
1

2

3

4

Fig. 1. Variations in mineral content obtained by XRD analysis of the surrounding soils and FMNs (bulk of all sizes) in non-paddy and paddy profiles of Wangling and Litang. S-Soil: surrounding soil, FMNs: iron-manganese nodules.

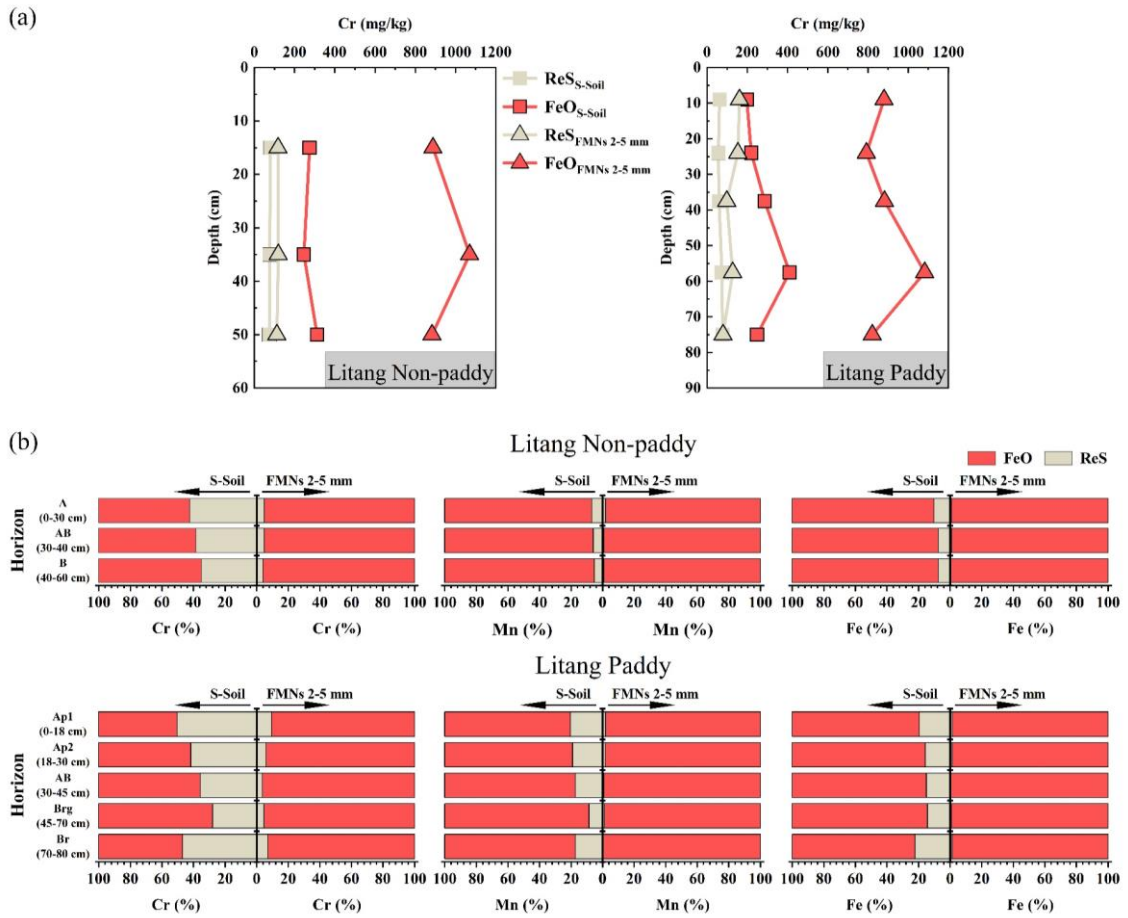


5

6 Fig. 2. Key element mass balance coefficients ($\tau_{X, Nb}$) in the non-paddy and paddy soil profiles of
 7 Wangling and Litang.

8

9



11

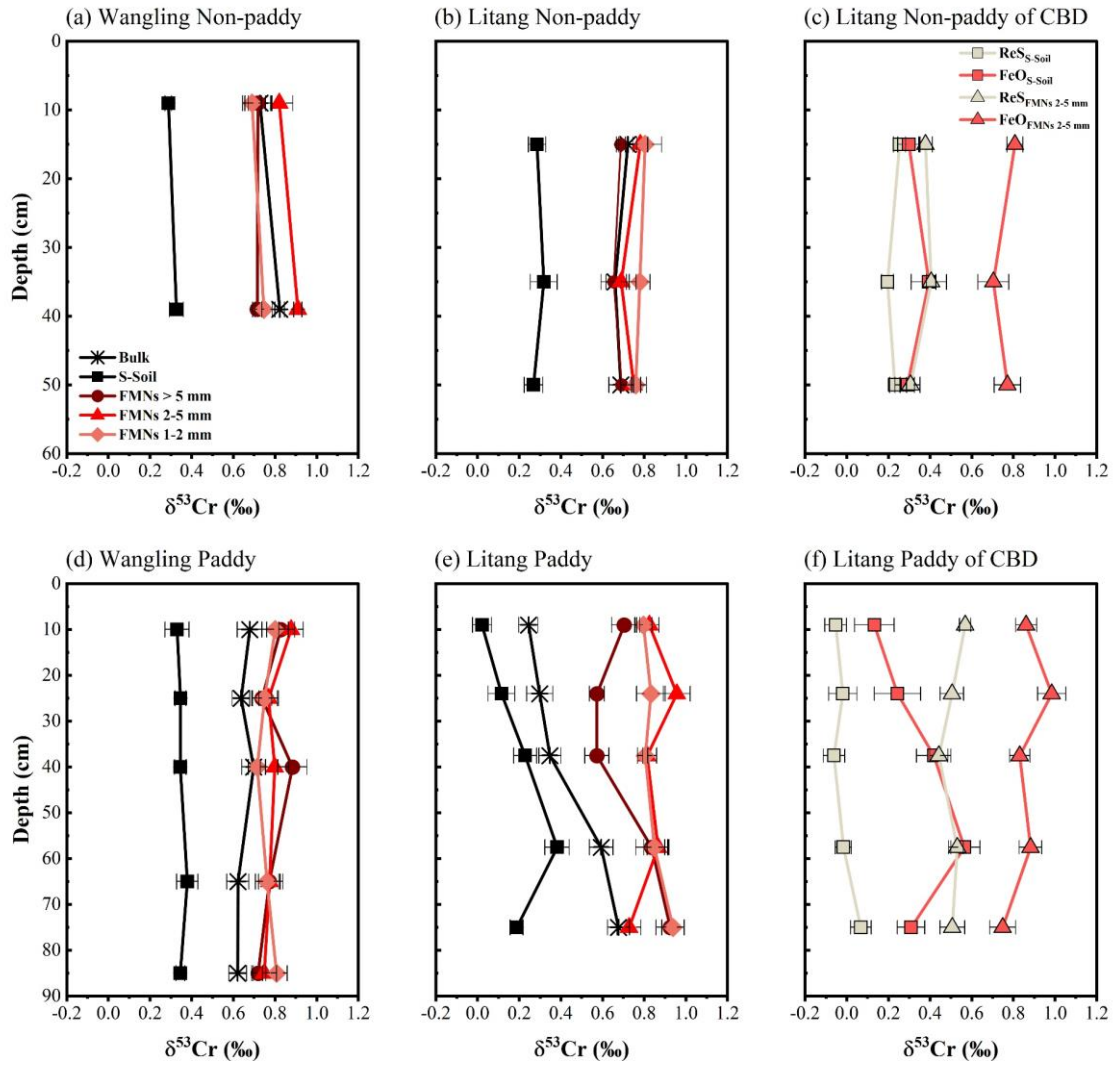
12

13

14

15

Fig. 3. CBD extraction results of (a) Cr concentrations of residue (ReS) and extract (FeO). (b) Relative distributions of Cr, Mn, and Fe between ReS and FeO in the surrounding soils and FMNs 2-5 mm in Litang soil profiles. S-Soil: surrounding soil, FMNs: iron-manganese nodules.



16

17

18

19

20

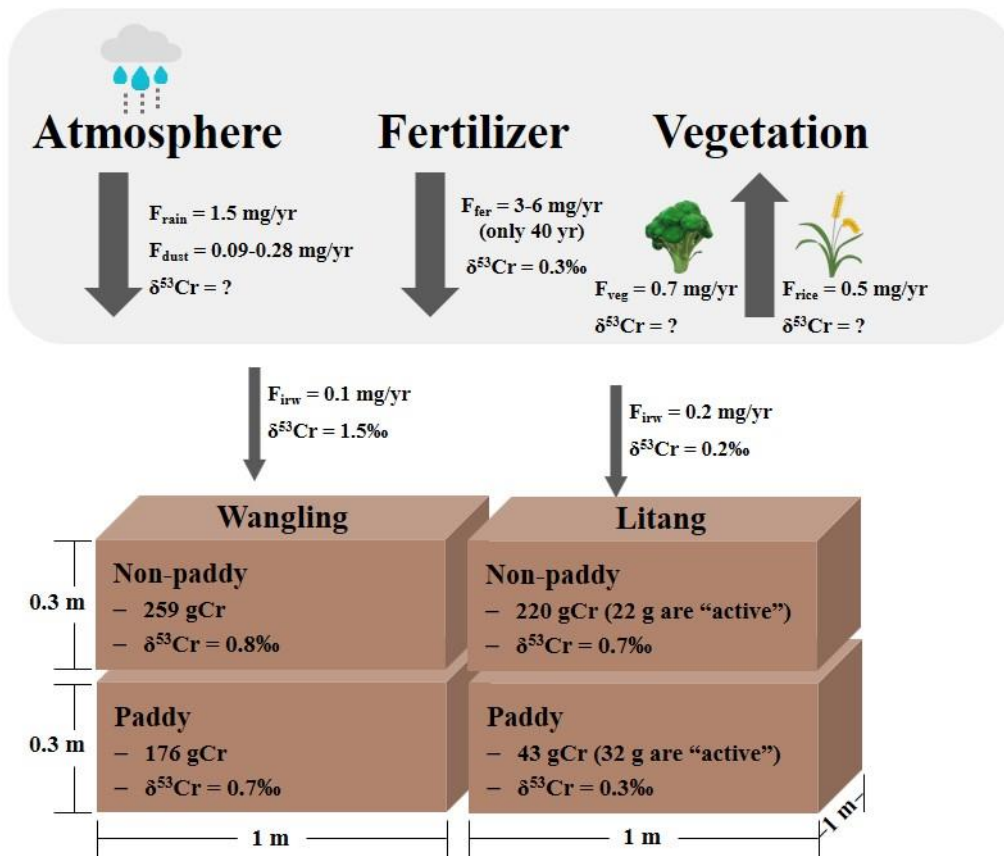
21

22

23

24

Fig. 4. Chromium isotope composition of samples in non-paddy and paddy profiles of Wangling, Litang before and after CBD extraction. Bulk: the bulk soil containing FMNs and surrounding soil, S-Soil: surrounding soil, FMNs: iron-manganese nodules, ReS and FeO: the residue and extract of CBD extraction, respectively. The error bars show the 2SD uncertainty derived from repeat measurements.



25

26

27

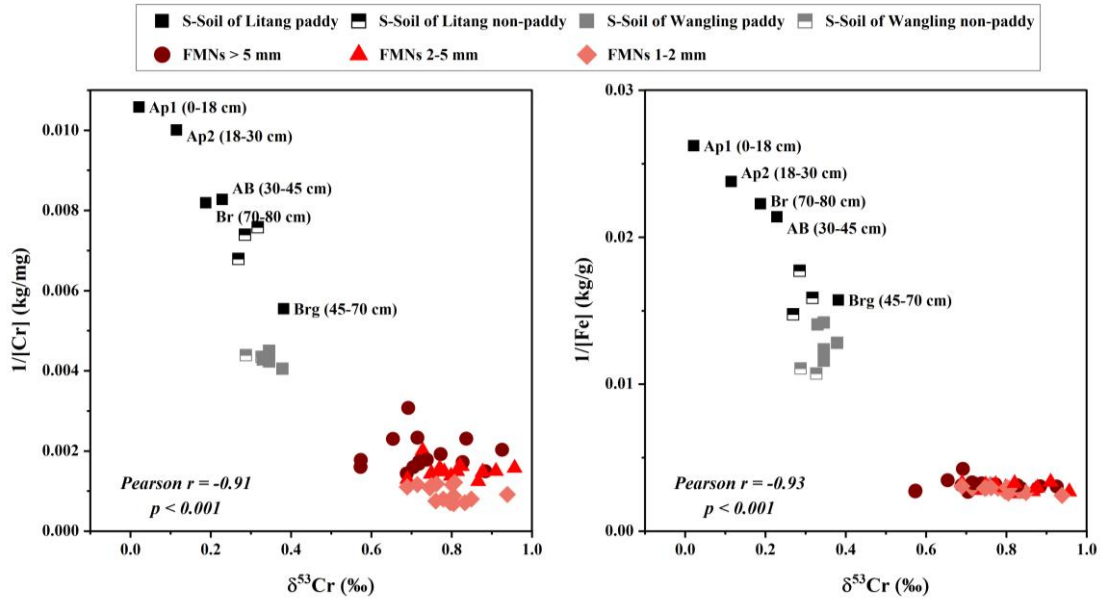
28

29

30

31

Fig. 5. Fluxes and stocks of Cr in the soil profiles of cultivated soil horizon in 0.3 m³ box at Wangling and Litang. F_{rain} : rainwater Cr flux, F_{dust} : human-derived dust Cr flux, F_{fer} : fertilizer Cr flux, F_{veg} : vegetable Cr flux in non-paddy field, F_{rice} : rice Cr flux in paddy field, F_{irw} : irrigation water Cr flux.



32

33

Fig. 6. $\delta^{53}Cr$ values vs $1/[Cr]$ and $1/[Fe]$ of the surrounding soils and FMNs of different sizes in

34

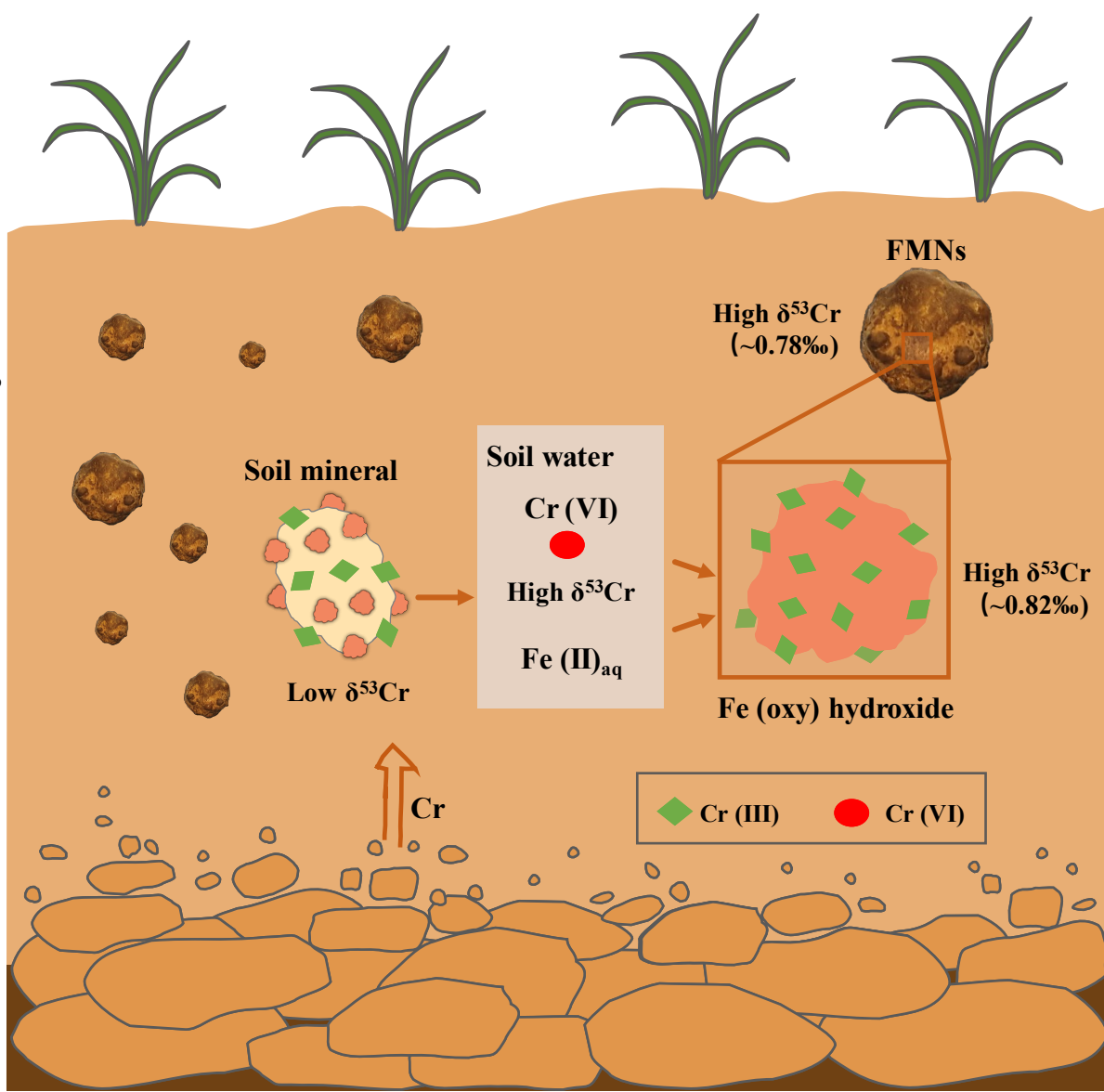
Wangling and Litang. S-Soil: surrounding soil, FMNs: iron-manganese nodules.

Environmental Implication

The existence of iron-manganese nodules (FMNs) in soil plays a crucial role in addressing soil pollution, especially when it comes to heavy metals. Our research demonstrates that FMNs found in red soils effectively capture and secure substantial quantities of the more easily movable and harmful form of chromium called Cr (VI), which also happens to be heavier in terms of its isotopic properties. This mainly happens because of the dominant presence of iron oxide in FMNs. This process significantly lowers the risk of chromium contamination in the red soils of the area we studied. Nevertheless, it's essential to stress the importance of further investigation into how stable these non-crystalline forms of iron oxide-bound chromium are in the surrounding soils over a long period.

Colluvium parent materials

S-Soil
 $\delta^{53}\text{Cr}$: 0.02-0.38‰



Highlights

- Cr isotope data in soil FMNs is reported for the first time.
- FMNs/Fe oxides capture isotopically heavy Cr (VI).
- Redox and not land use drives Fe oxide formation and FMNs, affecting Cr mobility.
- Cr in red soil profiles mainly results from colluvium mineral weathering.